





A transition to white LED increases ecological impacts of nocturnal illumination on aquatic primary producers in a lowland agricultural drainage ditch

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2 primary producers in a lowland agricultural drainage ditch

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20 **Abstract**

21 The increasing use of artificial light at night (ALAN) has led to exposure of freshwater
22 ecosystems to light pollution worldwide. Simultaneously, the spectral composition of
23 nocturnal illumination is changing, following the current shift in outdoor lighting technologies
24 from traditional light sources to light emitting diodes (LED). LEDs emit broad-spectrum
25 white light, with a significant amount of photosynthetically active radiation, and typically a
26 high content of blue light that regulates circadian rhythms in many organisms. While effects
27 of the shift to LED have been investigated in nocturnal animals, its impact on primary
28 producers is unknown. We performed three field experiments in a lowland agricultural
29 drainage ditch to assess the impacts of a transition from high-pressure sodium (HPS) to white
30 LED illumination (color temperature 4000K) on primary producers in periphyton. In all
31 experiments, we compared biomass and pigment composition of periphyton grown under a
32 natural light regime to that of periphyton exposed to nocturnal HPS or, consecutively, LED
33 light of intensities commonly found in urban waters (approximately 20 lux). Periphyton was
34 collected in time series (1 - 13 weeks). We found no effect of HPS light on periphyton
35 biomass; however, following a shift to LED the biomass decreased up to 62%. Neither light
36 source had a substantial effect on pigment composition. The contrasting effects of the two
37 light sources on biomass may be explained by differences in their spectral composition, and in
38 particular the blue content. Our results suggest that spectral composition of the light source
39 plays a role in determining the impacts of ALAN on periphyton and that the ongoing
40 transition to LED may increase the ecological impacts of artificial lighting on aquatic primary
41 producers. Reduced biomass in the base of the food web can impact ecosystem functions such
42 as productivity and food supply for higher trophic levels in nocturnally-lit ecosystems.

43

44 **Keywords:** artificial light at night; biofilm; light pollution; periphyton; urban stressor

45

46 **Capsule:** A switch from nocturnal high-pressure sodium to white LED illumination decreased
47 the biomass of periphyton in a field study; therefore a transition to LEDs in outdoor lighting
48 may increase ecological effects of light pollution in illuminated waters.

49

50 **Introduction**

51 The use of artificial light at night (ALAN) has rapidly increased in recent decades,
52 leading to light pollution – an increase of nocturnal light above natural levels, and a disruption
53 of the natural light/dark regime in many areas worldwide, with negative impacts on the
54 environment. Almost 25% of the world’s nonpolar land surface experiences light pollution
55 (Falchi *et al.*, 2016). This widespread environmental alteration contributes to global change
56 (Kyba *et al.*, 2017a), and raises concerns about the potentially adverse effects on organisms
57 and processes in illuminated ecosystems (Hölker *et al.*, 2010a, Rich & Longcore, 2006).
58 Along with an increase in nocturnal light levels, a global shift in outdoor lighting technologies
59 from narrow-spectrum, e.g. yellow high pressure sodium (HPS), to broad-spectrum, white
60 light-emitting diode (LED) lamps is taking place, driving a spectral shift in the nightscape
61 (Kyba *et al.*, 2017a). Replacement of traditional lighting technologies by energy efficient
62 LEDs is being implemented worldwide to decrease CO₂ emissions, environmental impacts,
63 energy consumption and lighting costs (Hölker *et al.* 2010b). However, there are concerns
64 about ecological impacts of LEDs (Davies *et al.*, 2017, Pawson & Bader, 2014, Ouyang *et al.*,
65 2017, Stone *et al.*, 2012) and the potentially negative consequences for human health
66 (American Medical Association, 2016) associated with blue light in nocturnal illumination,
67 which is found in commonly used white LEDs. The shift to LED lighting is expected to
68 increase the ecological impacts of ALAN for many organisms (Gaston *et al* 2012), but this
69 hypothesis has only been experimentally tested for terrestrial animals (e.g. bats, Lewanzik *et*
70 *al.* 2017, Rowse *et al.* 2016).

71 A substantial fraction of freshwater ecosystems is increasingly exposed to light
72 pollution worldwide. Estimates are that 80% of the human population lives in light-polluted
73 areas (Falchi *et al.*, 2016), more than half of the total population being situated within 3 km of
74 a body of freshwater (Kummu *et al.* 2011). With increasing population densities settling
75 closer to freshwaters (Ceola *et al.* 2015), these ecosystems experience drastic environmental
76 changes, including changes in light regime due to ALAN. Impacts of light pollution on
77 freshwaters, as on aquatic ecosystems in general are, however, substantially less studied
78 compared to terrestrial ecosystems (Perkin *et al.* 2011, Gaston *et al.* 2014). Furthermore, the
79 majority of studies on freshwaters have focused on invertebrates (e.g. Manfrin *et al.*, 2017,
80 Moore *et al.* 2008, Perkin *et al.*, 2014a, 2014b, Thomas *et al.* 2016) and fish (e.g. Brüning *et*
81 *al.*, 2015, 2018, Foster *et al.* 2016, Newman *et al.* 2015, Riley *et al.*, 2012, 2015), while
82 impacts of ALAN on microorganisms and primary producers, are still scarcely studied (e.g.
83 Hölker *et al.* 2015, Grubisic *et al.* 2017).

84 Primary producers use light as a source of energy for photosynthesis and as a source
85 of information about the external environment (Hegemann *et al.*, 2001). Intensity, spectral
86 composition, timing, and duration of light all influence photosynthesis (Falkowski &
87 LaRoche, 1991, Fortunato *et al.*, 2015), and ALAN may stimulate photosynthesis at times
88 when it would not naturally occur (Aube *et al.*, 2013). To contribute to net photosynthesis,
89 however, ALAN has to provide enough energy to activate the photosynthetic machinery and
90 to reach the photosynthetic compensation point. The intensity of illumination provided by
91 outdoor lighting (typically < 0.002 % of daytime light levels) is often considered to be too
92 low to stimulate photosynthetic production, except in directly illuminated areas (Gaston *et al.*,
93 2013). Theoretically, photosynthesis can occur at light levels slightly higher than moonlight,
94 i.e. $0.1 \mu\text{mol m}^{-2} \text{s}^{-1}$, approx. 0.5 – 7.4 lux (compared to the maximum moon light of 0.005
95 $\mu\text{mol m}^{-2} \text{s}^{-1}$, or approx. 0.3 lux, Kyba *et al.* 2017b) (Raven & Cockell, 2006), but the
96 minimum thresholds in a real-world setting are not well known. The light utilized in

97 photosynthesis, photosynthetically active radiation (PAR), ranges from 400 to 700 nm, but
98 blue (450 - 475 nm) and red light (630 - 675 nm) are utilized most efficiently. A light source
99 with strong emission in these spectral regions is thus more likely to stimulate photosynthesis
100 (Aube *et al.*, 2013). Furthermore, as light also provides information about the external
101 environment, changes in light intensity and quality are detected by photoreceptors and
102 conveyed for entrainment of the circadian clock that synchronizes internal physiological
103 processes with the external light/dark cycle (Fortunato *et al.*, 2015). In particular, blue light
104 perceived by cryptochromes and other flavin-containing is responsible for entrainment of the
105 circadian clock, which is crucial for photosynthesis, growth, and survival of algae (Fortunato
106 *et al.*, 2015, Dodd *et al.*, 2005). Disruption of natural light/dark cycles by ALAN, especially
107 by light sources with high blue content, may disrupt circadian regulation. White LED light,
108 with PAR efficiency at least twice than that of HPS (80-100% compared to maximum 40% of
109 HPS) (Darko *et al.*, 2014), and typically a high content of blue light in the emission spectra, is
110 therefore expected to have stronger impacts on aquatic primary producers compared to yellow
111 HPS light.

112 A few field and laboratory studies have investigated effects of ALAN on aquatic
113 primary producers, but the comparison of effects reported in these studies is confounded by
114 different light sources, light intensities, and studied experimental systems. Poulin *et al.* (2014)
115 found that short-term exposure to low-intensity HPS light (6.6 lux, approx. $0.08 \mu\text{mol m}^{-2} \text{s}^{-1}$)
116 affected several photophysiological processes in cyanobacterial cultures (*Microcystis*
117 *aeruginosa*), decreasing photosynthetic efficiency but not affecting growth. In microbial
118 sediment communities, one year of similar light exposure in the field (HPS light, 6.8 – 8.5
119 lux, approx. $0.09 \mu\text{mol m}^{-2} \text{s}^{-1}$) was found to alter community composition, increasing the
120 relative abundance of diatoms and cyanobacteria (Hölker *et al.* 2015). When the same
121 sediments were incubated in the laboratory under higher intensity white LED light (71 lux,
122 approx. $1.3 \mu\text{mol m}^{-2} \text{s}^{-1}$, color temperature 6300 K), nocturnal photosynthesis was stimulated

123 (Hölker *et al.*, 2015). In contrast, nocturnal illumination by white LEDs of intermediate
124 intensity (20 lux, approx. $0.31 \mu\text{mol m}^{-2} \text{s}^{-1}$, 3000K) decreased the biomass of periphyton in a
125 sub-alpine stream in a field study, and altered the proportions of diatoms and cyanobacteria
126 with the contrasting patterns (Grubisic *et al.* 2017, 2018). The role of spectral composition of
127 ALAN in its impact remains unclear, therefore whether a spectral shift in the nightscape can
128 affect aquatic primary producers is unknown. Whether ALAN affects only light-naïve
129 communities, or also those from areas that have previously experienced light pollution, is not
130 known.

131 We performed three experiments in a shallow, lowland agricultural ditch to assess the
132 effects of a shift from traditional HPS to white LED lighting, on periphyton. To mimic the
133 shift in outdoor lighting technologies, the HPS experiments were performed in 2014 (in
134 summer and winter), and the LED experiment was conducted in winter 2015. Periphyton was
135 collected in time series (1 - 13 weeks). In each experiment, the biomass and pigment
136 composition of periphyton exposed to nocturnal illumination were compared with those of
137 periphyton grown under a natural light regime. We expected yellow light from HPS lamps to
138 have a weak impact on periphyton, due to the mismatch of its spectral composition with the
139 sensitivity of photosynthetic pigments and its low blue content. We expected white LED light
140 to have a strong impact on primary producers, by stimulating photosynthesis through a greater
141 supply of PAR, and/or by disrupting circadian regulation because of high blue content.
142 Therefore, a shift from HPS to LED would increase impacts of ALAN on periphyton. As
143 periphyton is formed by a mixture of species that differ in their sensitivity to light and
144 photosynthetic optima (Jeffrey *et al.*, 1997), we expected ALAN to differently affect
145 individual taxa. This would drive a change in periphyton community composition that would
146 be reflected in an altered pigment composition.

147 **Material and methods**

148 *Sampling sites*

149 The experiments were conducted in 2014 and 2015 in a lowland freshwater system
150 within the rural area of Westhavelland Nature Park, located ca. 70 km northwest of Berlin, in
151 Brandenburg, Germany. The park has little artificial nocturnal illumination and has been
152 classified as a “Dark-Sky Reserve” by the International Dark-Sky Association (IDA). Three
153 experimental sites (Holzhauer *et al.* 2015) were chosen along an approx. 5m wide agricultural
154 drainage ditch (Fig. 1). The ditch is characterized by low flow velocity and soft sediment with
155 a mean annual depth of 50 cm. The treatment (Lit) site contained commercial streetlights, and
156 was designed to mimic street lighting conditions of urban areas. Illumination started in
157 summer 2012 and the lights were switched on at the beginning of the civil twilight and off at
158 dawn using an automatic time switch. The luminaires were equipped with 70 W HPS lamps
159 (VIALOX NAV-T Super 4Y, yellow 2000 K, Osram, Munich, Germany) from 2012 until
160 2015. In the summer of 2015 these were replaced with 51W white LED lamps (TECEO 1, 32
161 LEDs, neutral white 4000K, Schröder, Brussels, Belgium) that produced light of comparable
162 intensity at the water surface (19.5 ± 6.4 lux) to that of HPS lamps (17.1 ± 1.7 lux), directly in
163 front of the lamps. Light intensity was measured with an ILT1700 underwater photometer
164 (International Light Technologies Inc., Peabody, Massachusetts, USA), and the spectral
165 composition was measured using a compact spectrometer (specbos 1211, JETI, Jena,
166 Germany). The two control sites (Fig. 1) experienced a natural light regime with minimum
167 nocturnal light levels of 0.002 ± 0.001 lux (mean and *SD*) for Control 1 (at 800m distance
168 from the Lit site), and 0.010 ± 0.010 lux for Control 2 (at 300m distance from the Lit site).
169 Light intensity was measured on clear nights during the new moon phase after astronomical
170 twilight.

171 *Experimental design and sampling procedures*

172 For collection of periphyton, we constructed floating frames that held substrates for
173 periphyton growth (Fig. 2a), submerged in the upper 15 cm of the water column in the middle
174 of the ditch (Fig 2b). As substrates, we used plastic DIN A4 transparencies (polypropylene
175 sheets with a slightly textured surface; PolyClearView, IBICO, GBC, Chicago, IL, USA), that
176 were cut into strips (2 cm x 15 cm) that held together as a comb (Fig. 2a). The transparencies
177 were fixed with metal clips to the floating frames and held in a vertical position facing the
178 shore and the lamps (Fig 2b). These transparencies have been found to support the growth of
179 periphyton communities similar to those growing on natural substrates such as macrophytes
180 and muddy, organic sediments (Brothers *et al.*, 2013), that were also characteristic for this
181 ditch. A total of 108 strips were deployed at each site. For each sampling, we randomly cut 4
182 replicate strips from transparencies at each site, avoiding strips with visibly clear paths in the
183 periphyton indicative of grazing by snails. Strips were individually placed in 50-ml plastic
184 screw vials filled with pre-filtered ditch water (Whatman GF/F glass-fiber filter, 0.7 μ m
185 nominal pore size, Whatman Ltd., Maidstone, UK) and stored on ice in darkness pending
186 processing in the laboratory within 5 hours.

187 Two experiments with HPS lights were performed, in summer and winter 2014 (see
188 Fig. 3 for details of the timeline). On 21 July and 28 October (Week 0), we deployed the
189 frames with the substrates, and collected four 1-L water samples from the middle of the ditch
190 at all three sites (see below). In summer, we collected strips with periphyton after 1, 2, 3 and 4
191 weeks of growth. In winter, the growth of periphyton was slower (likely due to low water
192 temperatures), and the sampling was delayed to ensure that there was enough biomass for
193 analysis. As a result, strips were collected after 5, 6, 7, 8 and 13 weeks of growth in winter.
194 On 17 July 2015, HPS street lamps were replaced with LED lamps, and the experiment with
195 LED lights was started following an initial acclimation period, on 11 November 2015 (Week
196 0, Fig. 3), when we installed new frames with the substrates and collected water samples.

197 Strips with periphyton were collected on a weekly basis from weeks 1 to 6, and again at week
198 13.

199 In all experiments, illumination of the treatment site started before periphyton
200 exposure experiments began (Fig. 3). We therefore assessed potential chemical and biological
201 differences between treatment and control sites at the beginning of each experiment by
202 characterizing the water chemistry and phytoplankton community, each time we deployed
203 experimental substrates. Water chemistry (carbon and nutrient levels, See Supplementary
204 Material Table S1) was analyzed following standard chemical procedures (Krausse *et al.*,
205 1983, Murphy & Riley, 1962, Strickland & Parsons, 1968, Wetzel & Likens, 1991). Biomass
206 and pigment composition of suspended phytoplankton that was available to serve as the
207 founder community for the development of periphyton on the strips was assessed as described
208 below (Table S2). Environmental parameters (temperature, pH, dissolved oxygen, and
209 conductivity) were measured in time series, for each sampling event using a WTW Multi
210 3430 (WTW, Weilheim, Germany) equipped with WTW SenTix 940 pH sensor, WTW FDO
211 925 oxygen sensor, and WTW TetraCon 925 conductivity sensor (Tables S3-S5).

212 *Laboratory procedures*

213 Strips were removed from the 50-ml transportation vials. Periphyton was brushed
214 from the strips with a toothbrush, and rinsed with pre-filtered ditch water (Whatman GF/F
215 glass-fiber filter) into a graduated cylinder, to which the water used as transportation medium
216 was also added to ensure that none of the biomass was lost. The total volume of the resulting
217 suspension and the length of the strip that carried periphyton were recorded. After vigorous
218 shaking, aliquots for determination of dry mass were concentrated on pre-combusted, pre-
219 weighed 25 mm GF/F filters by vacuum filtration. Filters were dried at 65°C until constant
220 mass was achieved, and re-weighed. Additional aliquots for pigment analysis were
221 concentrated on filters and stored in 2-mL safety reaction vessels at -20°C pending analysis
222 by high-performance liquid chromatography (HPLC) (Waters, Millford, MA, U.S.A). These

223 filters were transferred to -80°C for a minimum of 48 hours to stimulate cell lysis, and
224 pigment composition was analyzed following the procedure described in Shatwell *et al.*
225 (2012). All manipulations were performed under dim light to avoid pigment degradation. Dry
226 mass and pigment composition of phytoplankton were analyzed from water samples following
227 the same procedure.

228 *Statistical analysis*

229 Each light source and season (HPS in summer and winter, LED in winter) were
230 analyzed separately. To test for effects of ALAN on periphyton biomass, we used generalized
231 least squares linear models (GLS) (Zuur *et al.*, 2009) as available in the nlme package
232 (Pinheiro *et al.*, 2015) for R (Version 3.1.3, R Development Core Team 2015). Site (“Lit”,
233 “Control 1”, “Control 2”), and time (weeks) were included as factors in the model. Time
234 series data were tested using the tseries package for R (Trapletti & Hornik, 2017). To account
235 for temporal correlation, the model included an auto-regressive correlation structure of order 1
236 (Zuur *et al.*, 2009). Biomass values were log- or square root-transformed when necessary to
237 improve distributional properties of the data. If variance heteroscedasticity was observed, the
238 factors (site or time) were used as variance covariates (Zuur *et al.*, 2009). Pairwise
239 comparisons were performed using the multcomp package for R (Hothorn *et al.*, 2008) with
240 Benjamini-Hochberg corrections for multiple comparisons. To test for differences in the
241 biomass of phytoplankton, GLS with site as factor was used, and post hoc comparisons were
242 performed using *t* tests with Benjamini-Hochberg corrections.

243 Pigment composition was used as a composite indicator of periphyton community
244 composition. Analysis of pigment composition, and especially pigment markers, is a
245 commonly used method for quantification of algal classes in mixed communities (Jeffrey
246 1997). Patterns in pigment concentrations in relation to ALAN were analyzed by applying
247 ordination analysis to all pigments identified in the samples. Pigment concentrations were
248 normalized to chl *a*, z-standardized, and analyzed using principal component analysis (PCA)

249 with functions from the *vegan* (Oksanen *et al.*, 2015), and *shape* (Soetaert, 2014) packages for
250 R. Pigment concentrations were log- or square root- transformed when necessary to meet the
251 assumptions of normal distribution. The scores of the first PCA component that accounted for
252 most of the variance in the data were statistically tested for effects of ALAN using the same
253 GLS models as mentioned above, i.e. for periphyton including site and time as fixed factors
254 and an auto-correlation structure; and for phytoplankton including site as fixed factor. Finally,
255 we performed a correlation analysis of the PCA component scores, to determine which
256 pigments were the drivers of variation in the data and the observed changes along the PCA
257 axes.

258 Environmental parameters (measured in time series) were compared across all sites
259 and three experiments by using GLS models with site as a factor, and an auto-regressive
260 moving average (ARMA) correlation structure that allowed for modelling time series at sites
261 for each experiment. When necessary, the variables were transformed and the model included
262 a correlation structure (site, season). Chemical parameters (measured only once and
263 transformed when needed) were compared among the sites for each experiment using one-
264 way analysis of variance (ANOVA) or Welsch test (when heteroscedascity was observed),
265 followed by tests for multiple comparisons with Bonferroni corrections.

266

267 **Results**

268 *Nocturnal HPS lighting*

269 In summer, the biomass of periphyton that developed on the strips increased with time
270 at all three sites (Fig. 4a. GLS: time effect $F_{1,41} = 76.53$, $p < 0.0001$). There were no
271 differences in biomass between the sites (GLS: site effect $F_{2,41} = 0.64$, $p = 0.53$), and no
272 significant interaction with time (GLS: site x time effect $F_{2,41} = 0.28$, $p = 0.75$).

273 Three chlorophyll pigments and five carotenoid pigments were identified in the
274 periphyton (Table S6). PC1 accounted for 54% of variance in the pigment composition and

275 was correlated with violaxanthin (Pearson's $r = -0.93$), lutein ($r = -0.92$), chl b ($r = -0.87$),
276 and neoxanthin ($r = -0.87$). PC1 clearly separated the Control 1 site from the two others (Fig.
277 5a), and PC1 scores indicated that periphyton pigment composition significantly changed over
278 time (GLS: time effect $F_{1,42} = 5.53$, $p = 0.02$), and across all sites (GLS: site effect $F_{2,42} =$
279 47.95 , $p < 0.0001$. t test: Lit to Control 1: $p < 0.0001$, Lit to Control 2: $p = 0.006$, Control 1 to
280 Control 2: $p < 0.0001$). There was no significant interaction (GLS: site x time effect $F_{2,42} =$
281 2.04 , $p = 0.15$).

282 In winter, the biomass of periphyton was highly variable (Fig. 4b) and changed
283 significantly over time (GLS: time effect $F_{1,66} = 20.15$, $p < 0.001$), and across sites (GLS: site
284 effect $F_{2,66} = 18.33$, $p < 0.0001$). The biomass at the Lit site did not differ from biomass at the
285 Control 2 site (t test: $p = 0.22$), and both were significantly higher than that of Control 1 (t
286 test: Lit to Control 1: $p < 0.0001$, Control 1 to Control 2: $p < 0.0001$). There was no significant
287 interaction (GLS: site x time $F_{2,66} = 1.81$, $p = 0.17$).

288 The same eight pigments that were identified in the summer (above) were also
289 identified in winter, along with one additional one (Table S7). PC1 accounted for 50% of the
290 variance in pigment composition and was correlated with diadinoxanthin (Pearson's $r = -$
291 0.94), and lutein ($r = -0.86$). Time-induced variation in pigment composition was visible
292 along PC1 (Fig. 5b), separating week 4 from the rest of the samples. PC1 scores indicated that
293 periphyton composition differed at all experimental sites (GLS: site effect $F_{2,54} = 5.93$, $p =$
294 0.005), and significantly changed over time (GLS: time effect $F_{1,54} = 15.22$, $p = 0.0003$). A
295 significant interaction (GLS: site x time effect $F_{2,54} = 6.55$, $p = 0.003$) indicated that
296 periphyton pigment composition changed in a different way across all sites over time.

297 Phytoplankton communities at the beginning of both HPS experiments did not show
298 patterns clearly linked to ALAN exposure: biomass did not differ between the two closer sites
299 (Lit and Control 2, apart by 300m from each other; mean and SD 4.1 ± 1.3 and 2.4 ± 1.2 in
300 summer, 39.6 ± 8.2 and 46.9 ± 12.1 in winter, respectively; compared to 15.2 ± 4.2 at Control

301 1 in summer and 67.5 ± 4.7 in winter; Table S8) despite the fact that the Lit site was exposed
302 to artificial nocturnal illumination for over one year. Phytoplankton differed in pigment
303 composition at all three sites (Tables S9-10).

304 *Nocturnal white LED lighting*

305 Following the transition of HPS to white LED lights, the biomass of periphyton was
306 significantly affected by site (Fig. 4c, GLS: site effect $F_{2,78} = 18.43$, $p < 0.0001$), but not by
307 time (GLS: time effect $F_{1,78} = 2.52$, $p = 0.12$). Starting from 3 weeks of growth until the end
308 of the experiment (13 weeks), the biomass at the Lit site was significantly lower compared to
309 Control 1 (t test: $p < 0.0001$), and Control 2 (t test: $p < 0.0001$) sites. The two control sites did
310 not significantly differ from each other (t test: $p = 0.06$). There was no significant interaction
311 (GLS: site x time effect $F_{2,78} = 1.08$, $p = 0.34$).

312 The same nine pigments that were identified in periphyton in previous winter
313 experiment (HPS) were also identified for LED experiment (Table S11). PC1 accounted for
314 68% of the variance in pigment composition (Fig. 5c), and correlated with chl c (Pearson's $r =$
315 0.94), violaxanthin ($r = 0.94$), fucoxanthin ($r = 0.93$), and diadinoxanthin ($r = 0.93$). Scores of
316 PC1 indicated different pigment composition of periphyton across all sites (GLS: site effect
317 $F_{1,75} = 30.84$, $p < 0.0001$. t test: Lit to Control 1 $p < 0.0001$, Lit to Control 2 $p < 0.0001$,
318 Control 1 to Control 2 $p < 0.0001$). There was no effect of time (GLS: time effect $F_{1,75} = 2.49$,
319 $p = 0.12$), and no significant interaction (GLS: site x time effect $F_{1,75} = 0.54$, $p = 0.58$).

320 Phytoplankton communities at the beginning of the LED experiment did not show
321 patterns that clearly linked to nocturnal light conditions: the biomass (mean and SD: Lit site
322 4.9 ± 1.1 , Control 1 17.0 ± 1.7 , Control 2 8.6 ± 1.8) and pigment composition differed from
323 each other at all three sites, Table S9-10.

324 **Discussion**

325 We found no effect of nocturnal illumination by HPS lights on periphyton in either
326 summer or winter, but following a replacement of the lights with white LEDs, the biomass of

327 periphyton that developed under a comparable nocturnal illumination (approx. 20 lux) in
328 winter was up to 62% lower compared to periphyton that experienced a natural light regime.
329 Neither light source had a substantial effect on periphyton community composition, as
330 indicated by the relative pigment concentrations. This study is the first to report the effects of
331 ALAN on periphyton in a lowland freshwater ecosystem, and partly confirm the findings
332 reported for periphyton from a sub-alpine stream (Grubisic *et al.* 2017), namely that exposure
333 to white LED light at night (approx. 20 lux) can decrease the biomass of freshwater
334 periphyton in a natural setting. The biomass decrease can also be significant in communities
335 from areas that experienced ALAN by HPS over multiple years, following a switch to LED
336 lights. The type of the light source, likely its spectral composition and in particular the blue
337 light component, seem to play a significant role in determining biological impacts of ALAN
338 on periphyton biomass.

339 ALAN replaces the dark phase in a natural light/dark cycle, creating an environment
340 with alternating phases of high-level natural light during the day and low-level artificial light
341 during the night. Since algae are exposed to cycles of light and dark in their natural
342 environment, and those have been stable over geological and evolutionary timescales, it is
343 generally assumed that dark periods are necessary for their optimal growth (Carvalho *et al.*,
344 2011). Some species indeed require a dark period to obtain maximum growth rate
345 (Zevenboom and Mur 1984). In many algal species DNA replication and cell division occur in
346 darkness, and some phases of cell division cycle might be sensitive to light (Edmunds 1988).
347 The interruption of a dark phase in a light/dark cycle by short periods of bright light was
348 found to negatively affect growth of several species of diatoms and cyanobacteria (Gibson &
349 Fitzsimons, 1991, Gibson & Fitzsimons, 1992, Zevenboom & Mur, 1984); although this has
350 not been observed for all investigated species. In plants and moss, altered light regime has
351 been shown to cause circadian stress, characterized by reduced photosynthetic efficiency and
352 altered expression of clock genes (Nitschke *et al.*, 2016, Okada *et al.*, 2009). Change in

353 photoperiod imposed by ALAN can have a major impact on the disruption of circadian
354 rhythms and physiology of plants (Kwak 2017). It is reasonable to assume that the disruption
355 of a natural light regime by ALAN is likely to cause a distortion in circadian regulation of
356 algae and cyanobacteria in periphyton as well.

357 Decreased biomass under similar levels of nocturnal white LED illumination was
358 reported in several studies on algae and plants. Exposure to LED (20 lux, approx. $0.31 \mu\text{mol}$
359 $\text{m}^{-2} \text{s}^{-1}$) decreased the biomass of periphyton from a sub-alpine stream by up to 57% over 3
360 weeks (Grubisic *et al.* 2017), and 2 weeks of exposure (30 lux, approx. $0.44 \mu\text{mol m}^{-2} \text{s}^{-1}$)
361 decreased the biomass of bean plants by 22% (Sanders *et al.* 2015). By contrast, white LED
362 light of a higher intensity (71 lux, approx. $1.3 \mu\text{mol m}^{-2} \text{s}^{-1}$) was found to stimulate nocturnal
363 photosynthesis in microbial communities (Hölker *et al.*, 2015). However, longer exposure
364 (over 2 years) to higher level of nocturnal HPS light (approx. 82 lux, $1 \mu\text{mol m}^{-2} \text{s}^{-1}$) was also
365 found to decrease biomass of plants by 44-56% (Kwak *et al.* 2018). This decrease was
366 associated with reduced photosynthetic efficiency and likely oxidative damage (Kwak *et al.*
367 2018).

368 The effect of ALAN on biomass of primary producers appears not to be linearly
369 related to its intensity. While low-level light fall below the photosynthetic compensation point
370 and have no effect on growth, and ALAN by LED of higher intensity can stimulate nocturnal
371 photosynthesis, LED of intermediate intensities can have unexpected negative effects on
372 biomass accrual. This may be caused by the interference of nocturnal illumination with the
373 entrainment of the circadian clock and thereby the physiological and developmental processes
374 that it regulates. The disruptive impact of nocturnal white LED light may be related to its blue
375 component to which algae and plants are highly sensitive (Fig. 6). Light emitted by HPS
376 lamps contains very little blue wavelengths; instead it is rich in yellow wavelengths that are
377 not efficiently absorbed by photosynthetic pigments and cryptochromes (Fig. 6). This
378 mismatch is likely to explain why no significant effect of nocturnal illumination by HPS on

379 periphyton communities was observed. For circadian disruption to occur under HPS
380 illumination, higher light intensities may need to be applied at night (Kwak *et al.* 2018).
381 Furthermore, low-level light that falls below the photosynthetic compensation point may still
382 keep photosynthetic machinery active at night, which might be energetically costly (Poulin *et*
383 *al.* 2014; Hölker *et al.* 2015).

384 ALAN by both HPS and LED illumination was previously reported to affect
385 community composition of aquatic microbial communities, which is to be expected given a
386 large variability in light optima and sensitivity between the taxa. Relative proportions of
387 diatoms and cyanobacteria were altered by both short-term exposure (2-4 weeks) to mid-
388 intensity LED illumination (20 lux, approx. $0.31 \mu\text{mol m}^{-2} \text{s}^{-1}$) (Grubisic *et al.* 2017, 2018),
389 and by long-term exposure (over 1 year) to low-level HPS illumination at night (6.8 – 8.5 lux,
390 approx. $0.09 \mu\text{mol m}^{-2} \text{s}^{-1}$) (Hölker *et al.* 2015). These effects were often season-dependent. In
391 our study, there was a large variation in pigment composition of periphyton in all
392 experiments, and no clear patterns could have been linked to nocturnal light conditions. It is
393 possible that the detection of effects of ALAN on periphyton community composition was
394 obscured by this variation. Even though pigment analysis can be a useful method for detection
395 of ALAN effects on periphyton in outdoor mesocosm experiments (e.g. Grubisic *et al.* 2018),
396 a method that provides better taxonomic resolution such as metabarcoding or microscopical
397 identification may be more suitable for the field studies performed in more variable
398 conditions.

399 Seasonal variation in environmental conditions, including light intensity and
400 photoperiod, is not only an important driver of the taxonomic succession of periphyton
401 communities, but also of their physiological acclimation to light (Biggs 1996, Falkowski and
402 LaRoche 1991). As likely a consequence, the sensitivity and response of periphyton to ALAN
403 varies with season (Grubisic *et al.* 2017, 2018). Here, effects of HPS were studied in two
404 seasons (and season strongly influenced periphyton biomass, GLS: $F_{1,117} = 11.63, p < 0.01$),

405 but effects of LED illumination were studied only in winter. The seasonality in potential
406 responses of periphyton to ALAN needs to be considered before drawing general conclusions
407 on effects of LED on periphyton.

408 Apart from light, nutrients and grazers are also strong determinants of periphyton
409 biomass (Stevenson, 1996). While environmental and chemical parameters showed seasonally
410 variable patterns, most parameters did not differ between experimental sites (Tables S12-13),
411 indicating similar conditions in the water across all sites. This was to be expected, given that
412 the sites were located in one continuous ditch system (Fig. 1). Some parameters differed
413 between the sites in the LED experiment (e.g. soluble reactive phosphorus, dissolved organic
414 carbon, some nitrogen fractions and dissolved silica, Table S13), however, these differences
415 are unlikely to explain the observed biomass decrease at the lit site, as the lowest levels of
416 almost all of these parameters were recorded at one of the two control sites. Dissolved silica,
417 although being lowest at the lit site, was of comparable levels to those measured in HPS
418 experiment in summer, when no decrease in periphyton biomass was observed. This indicates
419 that the biomass decrease observed at lit site under LED illumination cannot be explained by
420 nutrient limitation. The grazers that were present in the system belonged to free-swimming
421 invertebrates such as zooplankton and mayflies, and large grazers such as snails and fish
422 (pers. obs.). Grazing activity was not measured in this study; however, strips that were grazed
423 by snails were excluded from the analyses, and the abundance of free-swimming grazers in
424 winter was very low based on observations from previous years (A. Manfrin, pers. comm.).
425 Effects of grazing on periphyton were thus expected to be low in winter and unlikely to
426 explain our results. Indirect effect of ALAN on periphyton through top-down control cannot,
427 however, be excluded, especially in summer when grazers may be present in higher densities.

428 The experimental infrastructure was installed in 2012 in a rural and ALAN-naïve area
429 (Holzhauer *et al.* 2015). By using this experimental setup we were able to separate effects of
430 ALAN from other co-occurring factors typical of urban areas (e.g. increased nutrient levels,

431 chemical pollution, surface runoffs and sedimentation). The experimental sites were under the
432 same management regime (e.g. grass-cutting, flow regulation) to minimize factors that might
433 confound the effects of ALAN, thereby allowing for their quantitative comparison. By
434 monitoring environmental variables we tried to capture potential effects of other important
435 factors on periphyton. It is possible that the effects of HPS light were obscured by natural
436 variability in other environmental factors, especially in the winter experiment when there was
437 a large variation in the data (Fig.4, Table S4). Given the patterns we observed in the
438 biological and environmental data at the beginning and during the course of each experiment,
439 it is likely that ALAN was the strongest driver of the decline observed in periphyton biomass
440 under LED light. As the use of LED illumination keeps increasing (Kyba *et al.* 2017a), its
441 impacts on aquatic primary producers may be substantial not only in newly-lit areas (Grubisic
442 *et al.* 2017) but also in areas that experienced light pollution by traditional lighting
443 technologies over longer times. An ongoing transition to LED illumination may thus increase
444 impacts of nocturnal illumination on periphyton.

445

446 **Conclusions**

447 Nocturnal low-level white LED illumination was previously reported to decrease the
448 biomass of periphyton in a sub-alpine stream in spring and autumn (Grubisic *et al.*, 2017).
449 Here, we found that the biomass decrease also occurred in winter, under a comparable LED
450 illumination in a strongly contrasting environment, a lowland freshwater ecosystem. The
451 reduced growth under nocturnal low-light white LED illumination (approx. 20 lux) may be a
452 general response of periphyton to LED with a significant blue content. A better mechanistic
453 understanding is needed to predict ecosystem consequences of different ALAN sources and
454 light intensities, including the potential evolutionary adaptations and interactions with co-
455 occurring factors in urban waters (see also Perkin *et al.* 2011). Periphyton is an important
456 ecosystem component in streams, ponds, wetlands, and clear shallow lakes where it forms the

457 base of the food web and participates in biochemical cycling. The ongoing transition to LEDs
458 may increase ecological impacts of nocturnal illumination because decreased periphyton
459 biomass under white LEDs may potentially decrease nutrient turnover and food supply for
460 higher trophic levels, lowering production in the illuminated ecosystems.

461

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473

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669 **Figure Captions**

670 **Figure 1.** Aerial view of the study area located in Westhavelland Nature Park in
671 Brandenburg, Germany. Three experimental sites along a drainage ditch: Lit, Control 1 (C1)
672 and Control 2 (C2). Map data provided by Google, Digital Globe, GeoBasis-DE/BKG.

673 **Figure 2.** Scheme of the floating frames that were used to hold the substrates for
674 periphyton growth (transparencies cut in strips) (a), in a vertical position below the water
675 surface in the middle of the ditch (b).

676 **Figure 3.** Timeline of the experimental manipulations in three experiments performed
677 in two consecutive years under high-pressure sodium (HPS) and light-emitting diode (LED)
678 nocturnal illumination.

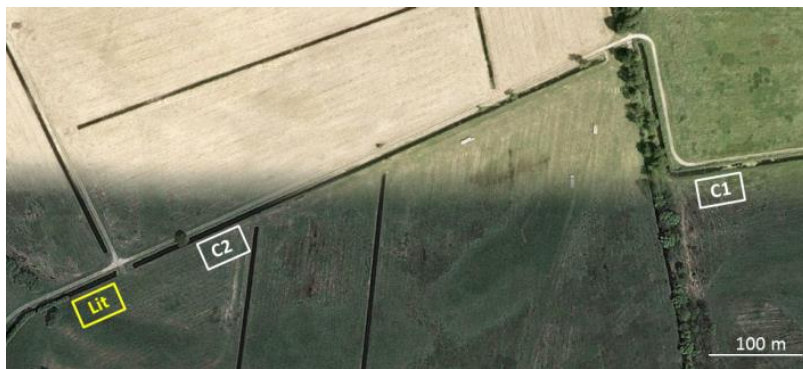
679 **Figure 4.** The biomass of periphyton (dry mass, mg cm^{-2} , mean and *SD*) developed at
680 the two control sites and at the Lit site, under high-pressure sodium lamps (HPS) in summer
681 (a, $n = 48$) and in winter (b, $n = 60$), or LED lamps in winter (c, $n = 84$).

682 **Figure 5.** Correlation biplots of principal component analysis (PCA) based on relative
683 pigment concentrations of periphyton, normalized to chlorophyll a, developed at the two
684 control sites and at the Lit site, under high-pressure sodium lamps (HPS) in summer (a, $n =$
685 48) and in winter (b, $n = 60$), or LED lamps (c, $n = 84$). Planes of the first two PC axes
686 explain 76% (a), 73% (b) and 89% (c) of the variation in the data. The pigments included in
687 the analysis were chlorophyll b (chl b), chlorophyll c (chl c), fucoxantin (fucox), violaxanthin
688 (violax), diadinoxanthin (diadinox), neoxanthin (neox), zeaxanthin (zeax), alloxanthin (allox),
689 and lutein. Arrows indicate correlation between original pigment variables and PC axes ($p <$
690 0.0001).

691 **Figure 6.** Spectral composition of the two light sources used in the study (high-
692 pressure sodium, HPS and light-emitting diodes, LED), spectral absorption curve of
693 chlorophyll a (based on Lohrenz *et al.*, 2003), and spectral sensitivity of cryptochromes
694 (based on Malhotra *et al.*, 1995).

695 **Figure 1.**

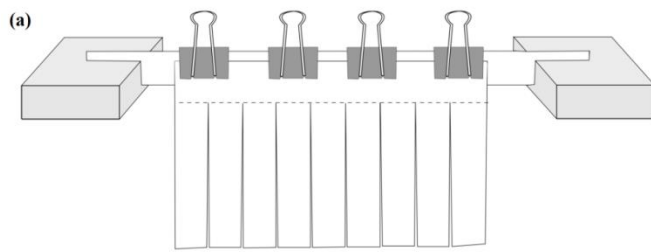
696



697

698 **Figure 2.**

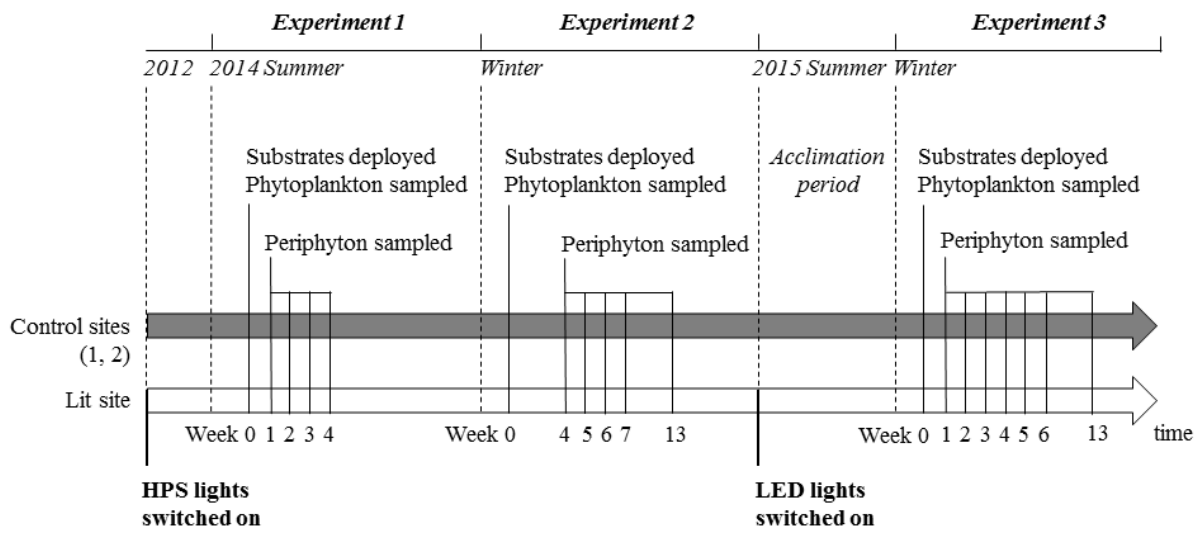
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700

701 **Figure 3.**

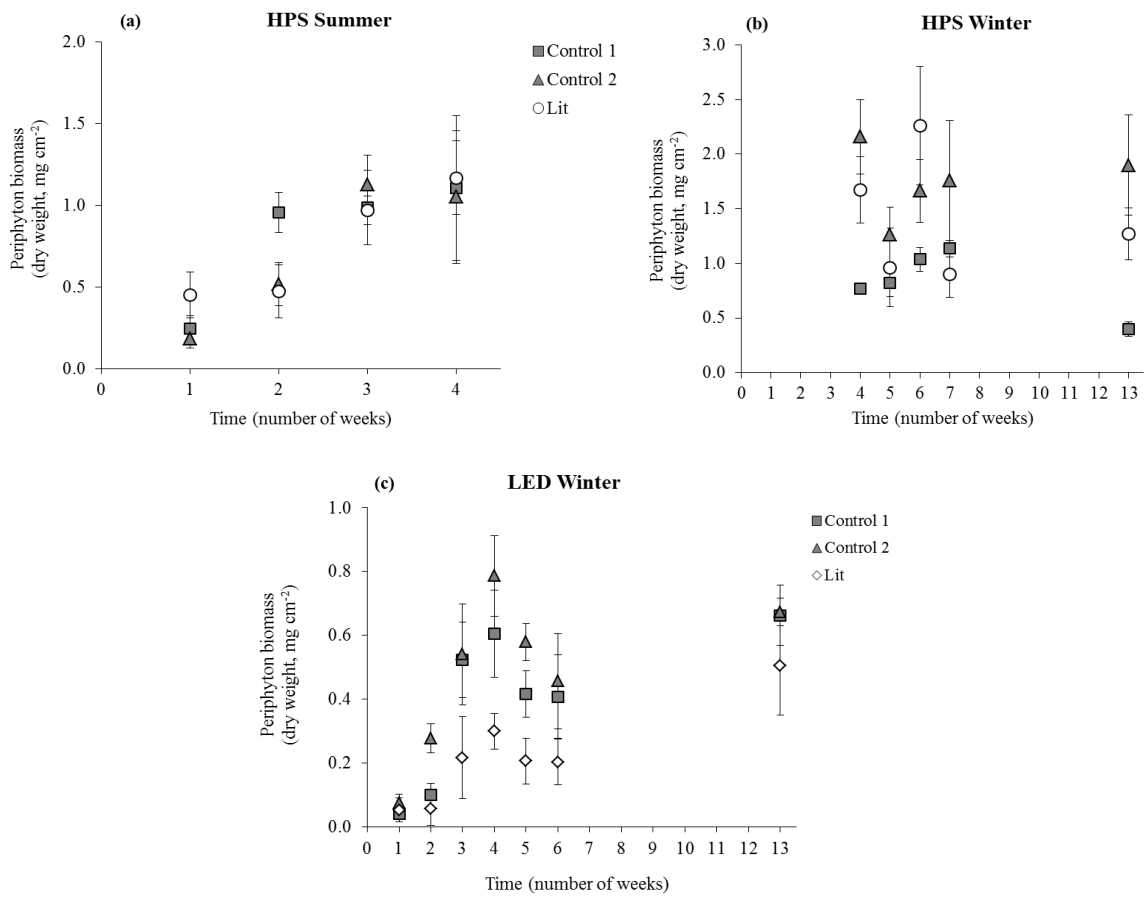
702



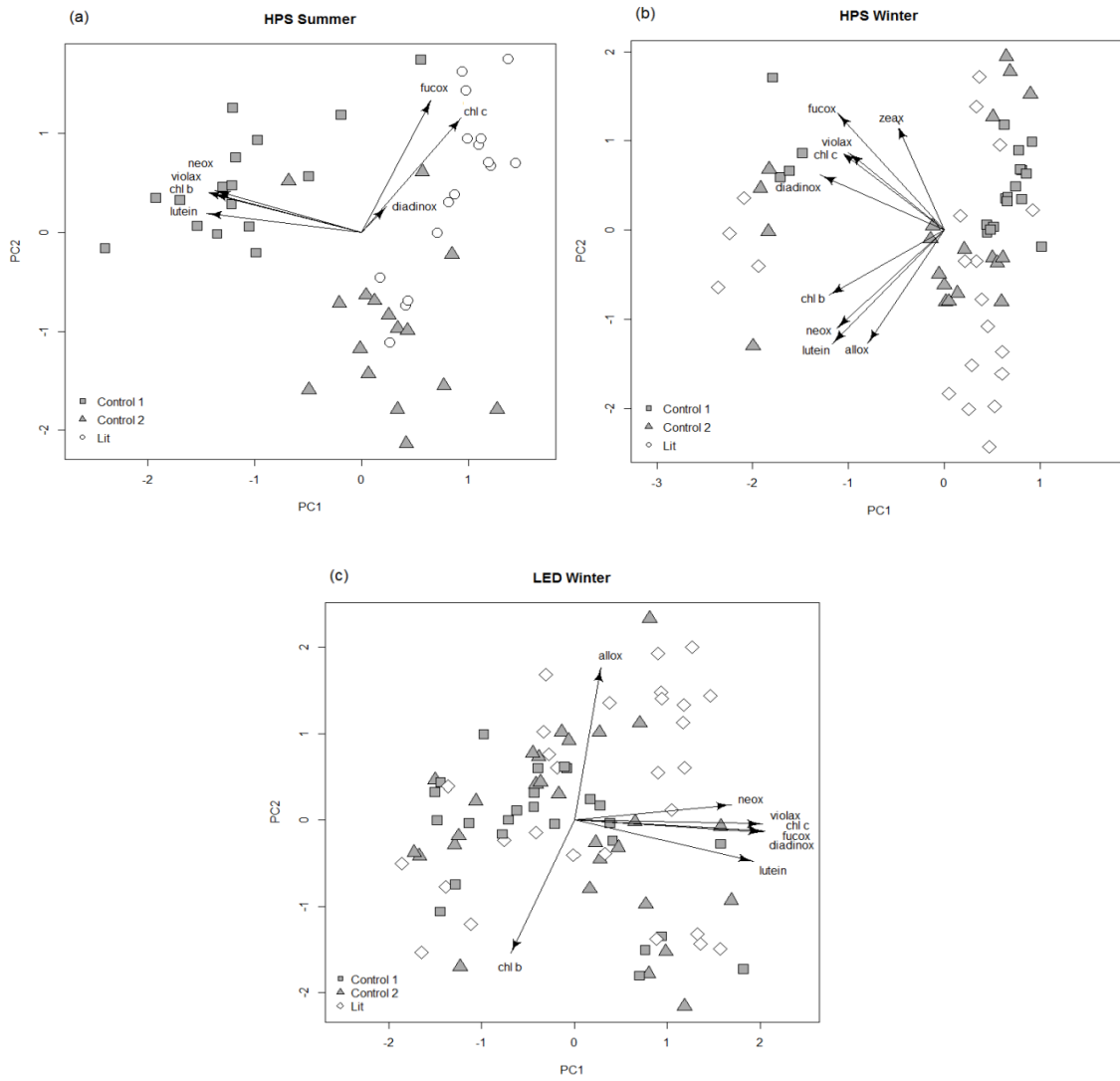
703

704 **Figure 4.**

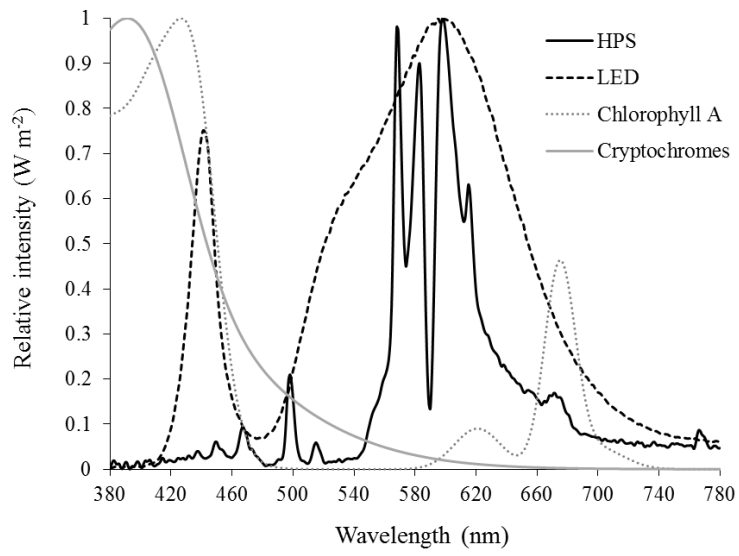
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710

Supplementary material for Grubisic *et al.* 2018 “*A transition to white LED increases ecological impacts of nocturnal illumination on aquatic primary producers in a lowland agricultural drainage ditch*”.

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Table S1. Chemical parameters ($\mu\text{g L}^{-1}$) measured at the three experimental sites at the beginning of each of the three experiments under high-pressure sodium (HPS) and light-emitting diode (LED) lights.

	HPS Summer			HPS Winter			LED Winter		
	Lit	Control 1	Control 2	Lit	Control 1	Control 2	Lit	Control 1	Control 2
Dissolved organic carbon	11.4	12.0	NA	10.3	10.0	NA	17.5	15.6	16.1
Dissolved nitrogen	0.70	0.79	NA	0.77	0.96	NA	1.75	1.56	1.63
Dissolved organic nitrogen	0.69	0.71	NA	0.59	0.64	NA	1.57	1.36	1.42
Nitrate-nitrogen	0.01	0.01	NA	0.01	0.01	NA	0.13	0.06	0.13
Nitrite-nitrogen	0.01	0.01	NA	0.01	0.01	NA	0.01	0.01	0.01
Ammonium-nitrogen	0.03	0.07	NA	0.17	0.31	NA	0.04	0.14	0.08
Soluble reactive phosphorus	24.7	63.0	NA	62.0	72.0	NA	15.0	40.0	11.0
Total phosphorus	94.0	132	NA	107.7	119.7	NA	63.0	75.3	64.6
Dissolved silica	3.27	6.12	NA	NA	NA	NA	4.83	6.07	5.27

Table S2. Pigment composition of phytoplankton ($\mu\text{g L}^{-1}$) developed at the three experimental sites under high-pressure sodium (HPS) and light-emitting diode (LED) lights.

	HPS summer			HPS winter			LED winter		
	Lit	Control 1	Control 2	Lit	Control 1	Control 2	Lit	Control 1	Control 2
Chlorophyll <i>a</i>	13.91	40.24	9.17	3.24	6.78	3.83	29.70	64.74	80.69
Chlorophyll <i>b</i>	2.97	10.59	1.25	n.d.	0.63	0.31	7.33	2.80	25.12
Chlorophyll <i>c</i>	0.29	0.68	0.33	n.d.	n.d.	n.d.	0.96	2.74	2.12
Fucoxanthin	0.64	2.17	0.46	0.25	0.65	0.18	2.72	10.93	6.90
Neoxanthin	0.17	0.78	0.05	n.d.	n.d.	n.d.	0.46	n.d.	1.26
Violaxanthin	0.33	0.66	0.09	n.d.	n.d.	n.d.	0.95	2.39	2.66
Diadinoxanthin	0.10	0.70	0.07	n.d.	0.21	n.d.	0.14	0.11	0.30
Alloxanthin	0.30	0.32	0.52	0.03	0.20	0.03	0.74	0.93	1.06
Lutein	0.58	2.01	0.21	n.d.	0.26	0.02	1.04	0.45	2.36
Zeaxanthin	0.08	0.48	0.08	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.

Table S3. Environmental parameters measured with the multi probes at the three experimental sites for the summer sampling under high-pressure sodium lamps (HPS).

HPS summer	Lit					Control 1					Control 2				
	0	1	2	3	4	0	1	2	3	4	0	1	2	3	4
time (weeks)															
Temperature (°C)	22.4	21.0	20.3	NA	16.8	26.7	23.4	21.7	NA	17.9	23.8	21.8	21.0	NA	17.0
Conductivity ($\mu\text{S cm}^{-1}$)	506	468	530	NA	424	530	479	505	NA	645	486	435	434	NA	406
Oxygen (mg L^{-1})	5.34	6.11	6.43	NA	6.40	5.77	4.51	8.82	NA	8.22	10.28	12.56	12.70	NA	10.40
Oxygen (%)	62.8	69.4	72.2	NA	67.6	73.1	53.6	105.6	NA	88.7	123.5	148.4	150.3	NA	110.0
pH	7.5	7.6	7.5	NA	7.4	7.6	7.7	7.9	NA	7.9	7.6	8.2	8.4	NA	8.0

Table S4. Environmental parameters measured with the multi probes at the three experimental sites for the winter sampling under high-pressure sodium lamps (HPS).

HPS winter	Lit						Control 1					
	0	4	5	6	7	13	0	4	5	6	7	13
time (weeks)	0	4	5	6	7	13	0	4	5	6	7	13
Temperature (°C)	11.3	3.5	3.2	3.0	2.9	3.1	11.6	3.4	2.4	2.0	2.5	3.2
Conductivity ($\mu\text{S cm}^{-1}$)	500	516	538	526	551	468	553	557	603	558	722	545
Oxygen (mg L^{-1})	1.46	4.10	2.45	4.77	9.70	10.84	2.44	6.73	4.91	7.18	12.57	12.55
Oxygen (%)	13.3	31.3	18.1	36.7	72.6	82.8	22.4	50.4	35.7	52.4	93.2	96.3
pH	7.1	7.5	7.3	7.6	7.8	7.7	7.2	7.7	7.5	7.6	7.9	8.0

	Control 2					
time (weeks)	0	4	5	6	7	13
Temperature (°C)	NA	3.5	2.2	2.6	2.8	3.0
Conductivity ($\mu\text{S cm}^{-1}$)	NA	506	535	532	558	485
Oxygen (mg L^{-1})	NA	5.06	2.68	5.20	11.00	11.00
Oxygen (%)	NA	38.0	19.3	38.5	82.6	84.1
pH	NA	7.6	7.2	7.4	7.8	7.7

Table S5. Environmental parameters measured with the probes at the three experimental sites for the winter sampling under light-emitting diode (LED) lights.

LED winter time (weeks)	Lit								Control 1							
	0	1	2	3	4	5	6	13	0	1	2	3	4	5	6	13
Temperature (°C)	10.6	10.4	2.0	5.5	7.6	4.7	7.8	5.9	10.8	10.4	2.8	5.6	7.5	4.7	7.9	5.8
Conductivity ($\mu\text{S cm}^{-1}$)	545	534	500	556	513	515	498	454	573	586	617	622	652	621	611	536
Oxygen (mg L^{-1})	6.90	3.15	5.60	8.15	8.60	8.60	7.60	13.45	7.25	4.60	5.65	9.45	9.80	9.10	6.80	11.80
Oxygen (%)	63.0	28.3	40.6	64.5	70.4	66.2	63.7	110.8	66.0	42.0	42.8	74.5	80.3	69.8	57.5	97.2
pH	7.6	7.3	7.4	7.5	7.2	7.5	7.5	7.8	7.6	7.3	7.6	7.6	7.5	7.5	7.5	7.8

	Control 2							
time (weeks)	0	1	2	3	4	5	6	13
Temperature (°C)	10.5	10.4	2.6	5.6	7.8	4.7	7.9	5.9
Conductivity ($\mu\text{S cm}^{-1}$)	540	540	549	567	561	513	516	467
Oxygen (mg L^{-1})	6.10	4.50	7.70	9.80	9.95	9.20	10.25	13.95
Oxygen (%)	54.7	40.5	59.2	78.0	82.5	70.9	86.2	114.9
pH	7.5	7.3	7.6	7.5	7.4	7.7	7.7	8.2

Table S6. Pigment composition of periphyton ($\mu\text{g cm}^{-1}$) developed at the three experimental sites over 1 to 4 weeks of substrate incubation for the summer sampling under high-pressure sodium lights (HPS).

HPS summer time (weeks)	Lit				Control 1				Control 2			
	1	2	3	4	1	2	3	4	1	2	3	4
Chlorophyll <i>a</i>	1.02	0.75	1.44	1.34	1.23	1.68	1.84	1.31	0.39	1.18	1.83	1.57
Chlorophyll <i>b</i>	0.18	0.11	0.11	0.17	0.41	0.67	0.64	0.41	0.08	0.19	0.25	0.35
Chlorophyll <i>c</i>	0.05	0.02	0.05	0.05	0.04	0.02	0.04	0.04	0.01	0.01	0.03	0.03
Fucoxanthin	0.12	0.05	0.18	0.15	0.13	0.11	0.16	0.07	0.03	0.04	0.11	0.11
Neoxanthin	0.01	0.01	0.02	0.00	0.02	0.05	0.04	0.04	0.00	0.01	0.02	0.02
Violaxanthin	0.01	0.01	0.02	0.02	0.03	0.05	0.05	0.03	0.01	0.02	0.03	0.03
Diadinoxanthin	0.02	0.01	0.03	0.03	0.01	0.01	0.02	0.02	0.05	0.01	0.02	0.02
Alloxanthin	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Lutein	0.02	0.03	0.05	0.05	0.05	0.13	0.12	0.08	0.01	0.04	0.07	0.06

Table S7. Pigment composition of periphyton ($\mu\text{g cm}^{-1}$) developed at the three experimental sites over 4 to 13 weeks of substrate incubation for the winter sampling under high-pressure sodium lights (HPS).

HPS winter time (weeks)	Lit					Control 1					Control 2				
	4	5	6	7	13	4	5	6	7	13	4	5	6	7	13
Chlorophyll <i>a</i>	6.44	2.62	4.03	3.56	5.68	2.54	1.28	2.21	4.49	2.18	6.54	2.84	4.18	4.60	6.99
Chlorophyll <i>b</i>	1.88	0.43	0.59	0.40	0.51	0.73	0.17	0.27	0.35	0.30	1.73	0.36	0.53	0.57	0.43
Chlorophyll <i>c</i>	0.91	0.13	0.23	0.27	0.29	0.50	0.10	0.17	0.38	0.16	0.95	0.21	0.34	0.33	0.44
Fucoxanthin	2.53	0.36	0.70	0.76	1.50	1.14	0.27	0.48	1.00	0.50	2.59	0.50	0.84	0.90	1.58
Neoxanthin	0.20	0.03	0.04	0.03	n.d.	0.05	n.d.	0.01	0.02	n.d.	0.15	n.d.	0.03	0.04	n.d.
Violaxanthin	0.16	0.03	n.d.	0.03	0.07	0.06	0.01	0.02	0.03	0.02	0.15	0.03	0.06	0.09	0.06
Diadinoxanthin	0.34	0.05	0.08	0.08	0.17	0.13	0.03	0.04	0.10	0.04	0.30	0.05	0.09	0.12	0.16
Alloxanthin	0.15	0.03	0.03	0.02	0.01	0.01	n.d.	n.d.	n.d.	n.d.	0.18	0.03	0.06	0.06	0.03
Lutein	0.38	0.08	0.11	0.07	0.06	0.08	0.02	0.03	0.04	0.03	0.30	0.05	0.09	0.10	0.06

Table S8. Statistical analysis of biomass of phytoplankton communities that were available as founder communities for development on periphyton at the three experimental sites (Control 1 (C1) and Control 2 (C2) and Lit) under high-pressure sodium (HPS) and light-emitting diode (LED) lights. Data analyzed using generalized least squares linear models with fixed factor Site. Table shows estimates for factor Site and post hoc comparisons performed using *t* test with Benjamini-Hochberg corrections. Asterisks indicate a significant effect.

		<i>F</i>	<i>df</i>	<i>p</i>		
HPS	Summer	25.25	1, 2	< 0.001*	<i>t</i> test:	Lit to C1 <i>p</i> < 0.001* Lit to C2 <i>p</i> = 0.06 C1 to C2 <i>p</i> < 0.001*
	Winter	7.48	1, 2	0.01*	<i>t</i> test:	Lit to C1 <i>p</i> = 0.01* Lit to C2 <i>p</i> = 0.29 C1 to C2 <i>p</i> = 0.04*
LED	Winter	62.48	1, 2	< 0.0001*	<i>t</i> test:	Lit to C1 <i>p</i> < 0.001* Lit to C2 <i>p</i> < 0.001* C1 to C2 <i>p</i> = 0.002*

Table S9. Statistical analysis of pigment composition of phytoplankton developed under high-pressure sodium (HPS) and light-emitting diode (LED) lights. Scores of the first axis (PC1) calculated with principal component analysis based on relative pigment concentrations were tested using generalized least squares linear models with fixed factor Site. Table shows estimates for factor Site and post hoc comparisons performed using *t* test with Benjamini-Hochberg corrections. Asterisk indicates a significant effect.

		<i>F</i>	<i>df</i>	<i>p</i>		
HPS	Summer	143.26	1, 2	< 0.0001*	<i>t</i> test:	Lit to C1 <i>p</i> = 0.0001* Lit to C2 <i>p</i> < 0.0001* C1 to C2 <i>p</i> < 0.0001*
	Winter	44.74	1, 2	< 0.001*	<i>t</i> test:	Lit to C1 <i>p</i> < 0.001* Lit to C2 <i>p</i> < 0.001* C1 to C2 <i>p</i> = 0.55
LED	Winter	13.56	1, 2	< 0.0001*	<i>t</i> test:	Lit to C1 <i>p</i> = 0.03* Lit to C2 <i>p</i> = 0.002* C1 to C2 <i>p</i> = 0.03*

Table S10. Variance in the data explained by the first axis (PC1) calculated from principal component analysis based on relative pigment concentrations of phytoplankton at the beginning of three experiments. Table shows pigment that correlated with PC1 for each sampling and Pearson's correlation coefficient.

		Explained variance (%)	Correlated pigments	Pearson's <i>r</i>
HPS	Summer	52	Lutein	- 0.99
			Chlorophyll <i>b</i>	- 0.93
			Neoxanthin	- 0.93
	Winter	52	Lutein	- 0.99
			Diadinoxanthin	- 0.96
LED	Winter	50	Alloxanthin	- 0.97
			Violaxanthin	-0.92

Table S11. Pigment composition of periphyton ($\mu\text{g cm}^{-1}$) developed at the three experimental sites over 1 to 13 weeks of substrate incubation for the winter sampling under light-emitting diode light (LED).

LED winter time (weeks)	Lit							Control 1							Control 2						
	1	2	3	4	5	6	13	1	2	3	4	5	6	13	1	2	3	4	5	6	13
Chlorophyll <i>a</i>	0.02	0.13	0.36	0.88	0.83	1.44	3.61	0.05	0.20	0.61	1.26	1.72	2.72	3.31	0.17	1.12	1.90	2.80	3.07	2.49	4.60
Chlorophyll <i>b</i>	n.d.	0.02	0.08	0.14	0.14	0.25	0.76	0.01	0.04	0.12	0.19	0.35	0.45	0.85	0.06	0.20	0.32	0.36	0.34	0.49	0.84
Chlorophyll <i>c</i>	n.d.	0.01	0.01	0.04	0.04	0.09	0.18	n.d.	0.01	0.03	0.07	0.10	0.19	0.15	n.d.	0.05	0.10	0.19	0.15	0.13	0.26
Fucoxanthin	n.d.	0.01	0.04	0.15	0.15	0.32	0.62	n.d.	0.02	0.09	0.25	0.35	0.71	0.53	0.01	0.13	0.27	0.65	0.62	0.56	0.89
Neoxanthin	n.d.	n.d.	n.d.	0.01	0.01	0.01	0.03	n.d.	n.d.	0.01	0.01	0.01	0.02	0.03	n.d.	0.01	0.02	0.02	0.02	0.09	0.03
Violaxanthin	n.d.	n.d.	0.01	0.01	0.01	0.02	0.03	n.d.	n.d.	0.01	0.02	0.02	0.03	0.03	n.d.	0.02	0.03	0.03	0.03	0.03	0.04
Diadinoxanthin	n.d.	n.d.	n.d.	0.01	0.01	0.03	0.06	n.d.	n.d.	0.01	0.02	0.03	0.05	0.04	n.d.	0.02	0.03	0.06	0.06	0.05	0.07
Alloxanthin	n.d.	n.d.	0.01	0.02	0.01	n.d.	n.d.	n.d.	0.01	0.02	0.01	0.01	n.d.	n.d.	n.d.	0.03	0.04	0.03	0.02	0.02	n.d.
Lutein	n.d.	n.d.	0.01	0.02	0.01	0.02	0.07	n.d.	n.d.	0.02	0.02	0.03	0.03	0.07	0.01	0.03	0.04	0.04	0.04	0.04	0.08

Table S12. Statistical analysis of chemical parameters (in $\mu\text{g L}^{-1}$) measured at the three experimental sites (Lit, Control 1 (C1) and Control 2 (C2)) at the beginning of each of the three experiments performed under high-pressure sodium (HPS) and light-emitting diode (LED) lights. Asterisks indicate a significant difference.

<i>Comparisons between Lit and C1 site</i>					<i>Comparisons between Lit, C1 and C2 sites</i>							
	<i>Parameter</i>	<i>F</i>	<i>df</i>	<i>p</i>		<i>Parameter</i>	<i>F</i>	<i>df</i>	<i>p</i>	<i>t test</i>		
HPS Summer	SRP	146.57	1,6	< 0.001*	LED Winter	SRP	5406.9	2,3	< 0.001*	Lit to C1 p = 0.006*		
	TP	2.3	1,5	0.18							Lit to C2 p < 0.001*	
	NO3-N	Equal values across fields									C1 to C2 p = 0.005*	
	NO2-N	Equal values across fields					TP	4.99	2,6	0.06		
	NH4-N	1,3	0,001				NO3-N	24.52	2,6	0.001*	Lit to C1 p = 0.003*	
	DN	1,6	0,04								Lit to C2 p = 1	
	DON	0.55	1,6	0.48							C1 to C2 p = 0.003*	
	DOC	3.73	1,6	0.1			NO2-N	Equal values across all fields				
	Dsi	190.5	1,6	< 0.001*			NH4-N	7.4	2,6	0.02	Lit to C1 p = 0.03*	
HPS Winter	SRP	1.33	1,6	0.29						Lit to C2 p = 0.54		
	TP	2.28	1,6	0.18		DN	2.79		0.14	C1 to C2 p = 0.18		
	NO3-N	Equal values across fields				DON	3.77	2,6	0.08			
	NO2-N	Equal values across fields				DOC	56.26	2,6	< 0.001*	Lit to C1 p < 0.001*		
	NH4-N	240.1	1,6	< 0.001*						Lit to C2 p < 0.001*		
	DN	34	1,6	0.001						C1 to C2 p = 0.13		
	DON	3.5	1,6	0.11		Dsi	17.32	2,6	0.003*	Lit to C1 p = 0.003*		
	DOC	0.79		1 0.37						Lit to C2 p = 0.22		
	Dsi	NA								C1 to C2 p = 0.03*		

Table S13. Statistical analysis of environmental parameters measured with the multi probes at all experimental sites (Lit, Control 1 (C1) and Control 2 (C2)) compared across all three experiments performed under under high-pressure sodium (HPS) and light-emitting diode (LED) lights. Data analyzed using general least square model with Bonferroni correction for multiple comparisons. Asterisks indicate a significant difference.

<i>Parameter</i>	<i>F</i>	<i>df</i>	<i>p</i>	<i>t test</i>	<i>F</i>	<i>df</i>	<i>p</i>	
Temperature (°C)	0.02	1,50	0.97					
Conductivity (µS cm ⁻¹)	8.9	2,50	< 0.001*	HPS summer	1.99	2,9	0.19	
				HPS winter	8.91	2,14	0.003*	Lit to C1 <i>p</i> < 0.001* Lit to C2 <i>p</i> = 0.74 C1 to C2 <i>p</i> = 0.001*
				LED winter	0.31	2,21	0.73	
Oxygen (mg L ⁻¹)	0.83	2,50	0.44					
Oxygen (%)	0.75	2,50	0.47					
pH	1.48	2,50	0.24					