



# Quantitative response of riverine benthic invertebrates to sediment grain size and shear stress

Stefan Lorenz  <https://orcid.org/0000-0002-2785-3404>, Christian Wolter  <https://orcid.org/0000-0002-2819-2900>

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6 **Authors:** Stefan Lorenz<sup>1,\*</sup> and Christian Wolter<sup>2</sup>

7 Stefan Lorenz: Leibniz-Institute of Freshwater Ecology and Inland Fisheries, Department  
8 Biology and Ecology of Fishes, Müggelseedamm 310, 12587 Berlin,  
9 Germany.

10 Julius Kühn-Institut, Federal Research Centre for Cultivated Plants,  
11 Institute for Ecological Chemistry, Plant Analysis and Stored Product  
12 Protection, Königin-Luise-Str. 19, 14195 Berlin, Germany

13 \* corresponding author: [stefan.lorenz@julius-kuehn.de](mailto:stefan.lorenz@julius-kuehn.de)

14 Christian Wolter: Leibniz-Institute of Freshwater Ecology and Inland Fisheries, Department  
15 Biology and Ecology of Fishes, Müggelseedamm 310, 12587 Berlin,  
16 Germany.

17

18 <sup>1</sup> [stefan.lorenz@julius-kuehn.de](mailto:stefan.lorenz@julius-kuehn.de), <sup>2</sup> [wolter@igb-berlin.de](mailto:wolter@igb-berlin.de)

19

## 20 **Abstract**

21 The most widespread pressure impacting river ecological status is the degradation of key  
22 hydromorphologic elements, such as sediment type and flow rate. However, almost nothing is  
23 known about the quantitative relationship between benthic invertebrate abundance and these  
24 elements. This synthesis compiles quantitative data on physical requirements and thresholds  
25 for invertebrates relative to two hydromorphologic factors: substrate size and hydraulic

26 energy (measured as shear stress). Both factors are commonly a focus of river rehabilitation.  
27 However, we found only limited literature data that we could use to identify invertebrate  
28 preferences (189 taxa). Preferred substrate sizes of all stream epifauna we examined varied  
29 between 0.05 and 400 mm and that they prefer shear stresses from 0.13 to 3.67 N m<sup>-2</sup>. There  
30 was no difference in variation of preferred conditions between the examined taxonomic  
31 levels. We suspect that taxa preferring hydraulic environments with shear stresses <0.64 N m<sup>-2</sup>  
32 are affected more by environmental factors than solely being constrained by substrate or  
33 hydraulic energy preferences. Such taxa might be useful as sensitive indicator species for  
34 evaluating stream integrity. Hence, to optimize restoration success for riverine biota,  
35 hydromorphological forces should be mitigated by manipulating habitat complexity in a way  
36 that it enhances intact ecological processes.

37

38 **Keywords:**

39 hydraulic preference, hydromorphology-biota interaction, indicator species, river integrity,  
40 substrate preference

41

42

43 ***Introduction***

44 Worldwide, hydromorphologic (the simultaneously operating processes of hydrology and  
45 morphology) and habitat alterations have been identified as the most significant pressure  
46 impacting the ecological integrity of riverine ecosystem, riverine biodiversity, and water  
47 quality (USEPA, 2009; Vaughan et al., 2009; EEA, 2012; Fehér et al., 2012; Elozegi &  
48 Sabater, 2013). An exponentially increasing number of river restorations have been attempted  
49 to enhance the hydromorphologic condition of riverine ecosystems (Bernhardt et al., 2005;  
50 Roni et al., 2005; Wolter, 2010; Palmer et al., 2010; Feld et al., 2011), but many of them have  
51 failed to reach their objectives. Restoration failure is commonly ascribed to the inappropriate  
52 spatial scale at which restoration activities are implemented (Sponseller et al., 2001; Kail &  
53 Hering, 2009), poor or neglected habitat enhancement during restoration (Miller et al., 2010;  
54 Hering et al., 2015), or the confounding impacts of multiple stresses at a variety of scales  
55 (Weigel et al., 2003; Robson & Mitchell 2010; Roni et al., 2008; Feld & Hering, 2007; Lorenz  
56 & Feld, 2013; Verdonschot et al., 2013). However, we currently lack a mechanistic  
57 understanding of how species respond to specific hydromorphologic conditions or how  
58 essential habitats or relevant bottlenecks are affected by hydromorphology (e.g., Wolter et al.,  
59 2004), thus leading to failure of some restoration projects.

60         Appropriate hydraulic flow and habitat structure are the two most significant  
61 parameters determining the ecological integrity of river epifauna, both of which are  
62 influenced by the interplay of river flow and channel morphology (Lorenz et al., 2016a).  
63 Dynamic flow is a hydraulic process that controls geomorphic patterns and variables and  
64 subsequently the composition of aquatic biota at various spatial and temporal scales (Frissell  
65 et al., 1986; Robson & Chester, 1999; Chessman et al., 2006; Milner et al., 2015). Feld et al.  
66 (2011) provided numerous conceptual models to illustrate the variety of potential mechanisms  
67 and interactions that are possible (at various spatial scales) among hydrologic regime,

68 geomorphology, biota, water chemistry, and suspended load. However, such multi-linkage  
69 models have rarely been used to predict interactions in any measurable or quantifiable way, at  
70 not least at a scale that could be used to inform rehabilitation planning. Notable exceptions  
71 include Janauer et al. (2010), who reported that species-rich and diverse communities of  
72 macrophytes might not tolerate flow velocities  $>0.3 \text{ m s}^{-1}$ . Similarly, Wolter and Arlinghaus  
73 (2003) determined that fish hatchlings require flows  $<0.1 \text{ m s}^{-1}$ , whereas juveniles cannot  
74 tolerate flows  $>0.5 \text{ m s}^{-1}$ . In contrast, Statzner et al. (1988) and Söhngen et al. (2008)  
75 determined that rheophilic invertebrates could tolerate flows from  $0.3$  to  $1.0 \text{ m s}^{-1}$ . However,  
76 such thresholds widely vary among higher taxa and even within genera (e.g.,  $<0.8 \text{ m s}^{-1}$  for  
77 gastropods,  $>1.5 \text{ m s}^{-1}$  for some dipterans, and  $>2.0 \text{ m s}^{-1}$  for some beetles (Statzner et al.,  
78 1988). Limnophilic species prefer substantially lower flows (e.g.,  $<0.2 \text{ m s}^{-1}$ ). Rheobiont  
79 macroinvertebrates that prefer flow velocities  $>1 \text{ m s}^{-1}$ , are rather rare (Söhngen et al., 2008).

80 Discharge (the volume of water discharged per unit time) and flow regime (both of  
81 which interact with valley and channel form) are extremely important in shaping  
82 hydromorphologic features and structures (Poff et al. 1997, Lorenz et al. 2016a). However,  
83 aquatic organisms typically do not respond to discharge directly, rather they typically respond  
84 to flow velocity and shear stress at their point of attachment. Both flow velocity and shear  
85 stress can limit habitat utilization by organisms. High flow rate and stream power regulate  
86 sediment transport and sorting. Hence, the array of substrate types formed by flow-induced  
87 sediment sorting processes are predictable and specific to a given type of hydraulic regime  
88 (Davis & Barmuta, 1989). Therefore, substrate characteristics may be useful for indicating  
89 hydromorphological integrity. Likewise, because specific taxa are associated with specific  
90 substrate types (which in turn indicate functioning fluvial processes), the taxa could be used to  
91 indicate hydromorphologic conditions.

92 Benthic invertebrates may be particularly useful as indicator taxa because their  
93 distribution is determined by primarily hydraulic and/or substrate conditions (Hart & Finelli,  
94 1999; Jowett, 2003; Gabel et al., 2012). Invertebrates are commonly used as water quality  
95 indicators in ecological assessments; under European environmental legislation, they are an  
96 obligatory biological quality component (Birk et al. 2012; Lyche-Solheim et al., 2013).  
97 However, little detailed information is available about the types of substrates stream  
98 invertebrates prefer or the shear stresses they can tolerate. For this reason, habitat preferences  
99 of stream invertebrates are commonly expressed in more qualitative terms provided in trait-  
100 databases (e.g., lithal and rheophilic) (Schmidt-Kloiber & Hering, 2015). In addition,  
101 substrate types and flow regimes are usually not very homogeneous in streams; rather, they  
102 are patchily distributed and occur at small spatial scales. This small-scale heterogeneity,  
103 combined with the high mobility of biota, may hamper the identification of indicator species  
104 that possess tolerances for specific hydromorphological conditions.

105 Some low-gradient river systems are naturally dominated by fine sediments (Downes  
106 et al., 2006) or have low oxygen concentrations (Sundermann et al., 2011a). In such systems,  
107 aquatic communities might tolerate sub-optimal substrate conditions if patches of limiting  
108 resources (such as food or oxygen) occur there (Downes et al., 2006; Robson et al., 1999).  
109 Additionally, substrate preference and small-scale distribution of invertebrates (such as  
110 shredder or collectors) are associated with the presence or absence of particular types of food  
111 resources (Culp et al., 1983; Williams & Moore, 1986). Substrate characteristics are known to  
112 influence food-web structure (Power et al., 2013) and determine the distribution of  
113 invertebrate larvae by influencing the types of habitats available for egg laying (Downes &  
114 Lancaster, 2010). However, invertebrate presence is not exclusively controlled by substrate  
115 characteristics; it is also influenced by how organisms interact with confounding  
116 environmental factors (Jowett, 2003). Unfortunately, it is difficult to determine to what degree

117 a species' presence is related to its hydraulic preferences, availability of food resources,  
118 oxygen demand ,or some combination thereof (e.g., Lorenz et al., 2016a), particularly because  
119 species-specific data on biological requirements are unavailable for most species (Tyler et al.,  
120 2012). Hence, river restoration needs also information if biological requirements from higher  
121 taxonomic levels may potentially be transferred to species level if species data are  
122 unavailable.

123         This study focuses on the physical thresholds of shear stress in rivers and evaluates the  
124 relationship to sediment size and distribution of riverine benthic invertebrate epifauna  
125 (hereafter called *invertebrates*), even though these variables form only a small but highly  
126 relevant component of hydromorphology (Williams & Hynes, 1974; Rabeni & Minshall,  
127 1977; Reice, 1980). Therefore, this study aims to develop a mechanistic understanding of the  
128 effects that key hydromorphologic variables have on invertebrate distributions to provide  
129 baseline information that can be used for planning successful river rehabilitations. We are  
130 well aware that invertebrate life histories and other factors, such as food preferences, also  
131 determine species-habitat relationships; however, each species reacts to dynamic flow and  
132 shear stress in its own way (e.g., suspension feeders obligatorily depend on flow) (Chester &  
133 Robson, 2011; Grown & Davis, 1994).

134         The main objectives of this synthesis were to (1) determine tolerance thresholds of  
135 physical forces limiting habitat suitability for riverine epibenthic species, (2) identify  
136 epibenthic species sensitive to variations across a wide range of hydromorphologic  
137 conditions, and (3) derive significant information if biological traits from higher taxonomic  
138 levels may potentially be transferred to species level.

139

## 140 **Methods**

141

142 Data collection

143 We acquired data from 12 studies (Online Resource 1) that used FST-  
144 (FließwasserStammTisch) hemisphere numbers to assess invertebrate hydraulic preferences  
145 (e.g. Dolédec et al., 2007; Mérigoux et al., 2009). FST hemisphere studies use a set of 24  
146 standard-sized hemispheres of identical size and texture, but each hemisphere varies in  
147 density (Statzner et al., 1991). Movement of the heaviest hemisphere at the stream bottom  
148 determines the shear stress acting at that location on the sediment surface. FST hemisphere  
149 numbers are converted to bed shear stress values according to Statzner et al. (1991). The shear  
150 stress measured at any given location can be used to determine the hydraulic preference of  
151 invertebrates inhabiting that particular hydraulic regime. Therefore, bed shear stress values  
152 can be used to predict suitable habitat for any invertebrate species across an array of potential  
153 fluvial environments (Gore, 1996; Möbes-Hansen & Waringer, 1998). Shear stress accounts  
154 for turbulence at near-bed surfaces generated by sediment roughness, which creates drag and  
155 lift forces acting on invertebrates inhabiting the substrate (Möbes-Hansen & Waringer, 1998;  
156 Mérigoux & Dolédec, 2004). Although the available FST hemisphere numbers were obtained  
157 for rivers and streams of various sizes, the hydraulic habitat preferences obtained seemed  
158 generally applicable as discussed by Lamouroux et al. (2012). We also explicitly included in  
159 our analysis taxa with low strength ( $r^2$ ) of the average taxa preference model (*sensu* Mérigoux  
160 et al., 2009). Unusual or unexpected hydraulic/substrate combinations resulting from such low  
161 strengths of taxa preference models might be indicative for taxa responses to particular  
162 ecological resource rather than to substrate type or flow condition in a stream (Kakouei et al.,  
163 2017). We reviewed additional information on flow velocity preferences for biota using the  
164 AQEM/STAR macroinvertebrates database (Schmidt-Kloiber & Hering, 2015;  
165 [www.freshwaterecology.info](http://www.freshwaterecology.info)) and additional scientific literature. Furthermore, we compiled



166 available data on the average shear stresses necessary to detach and dislodge invertebrates  
167 based on a Web of Science search.

168 We reviewed data from three studies that provided information on sediment habitat  
169 requirements of a large number of invertebrate species. Two studies, Tolkamp (1982) and  
170 Singh et al. (2010), provided an index of representation (IR), which provided habitat  
171 preferences or avoidance of habitats relative to specific grain size fractions (expressed as Phi  
172 indices) for a range of species. [The Phi index is the negative binary logarithm of particle size  
173 (Krumbein & Sloss, 1963).] Hence, we associated IR values for each taxon with the Phi  
174 indices provided in these studies and converted the indices to grain size fractions. The third  
175 study, Schröder et al. (2013), provided substrate preferences based on multi-level pattern  
176 analysis.

177

#### 178 Data analysis

179 To investigate whether the variation in hydraulic and substrate preferences differed among  
180 taxonomic levels (family, genus, and species) identified for each preference group, we first  
181 compared the interquartile ranges (IQR) of each genera's hydraulic and substrate preferences  
182 to the IQR of the family for the genus. Due to the lack of enough available hydraulic and  
183 substrate preference data for both taxonomic levels, family and genus, we limited our  
184 comparisons to 13 families with 16 genera for substrate preferences and to 19 families with 18  
185 genera for hydraulic preferences. Then we applied one-sided Wilcoxon tests (*wilcox.test*  
186 function, stat package, R statistical language, version 3.2.4) (R Core Team, 2015) to the IQRs  
187 to assess whether preferences were more variable within families than within genera.

188 To identify species whose presence were not likely determined primarily by  
189 hydromorphic differences (substrate type or hydraulic preference), we plotted shear stress  
190 relative to substrate classes (from 0 to 11) (*sensu* Singh et al. (2010), see Table 1 for

191 assignment of substrate class to grain size). Due to the very low number of species for which  
192 we could obtain data for both hydraulic and substrate preference data, we had to pool most  
193 species with other members of their genus, tribe or subfamily (e.g., dipterans and particularly  
194 chironomids) to obtain a large enough sample size ( $n \geq 42$ ). Some species show a large range  
195 of preferred grain sizes, e.g. the taxon *Rhithrogena* sp. shows a range of preferred grain sizes  
196 from 0.6 to 400 mm (Online Resource 3), which corresponds to a range of substrate classes  
197 from 0 to 8. We are aware that using mean values for a variety of related taxa will mask the  
198 narrow substrate preferences of specialist species and possibly artificially exaggerate variation  
199 in more generalist taxa. However, the lack of sufficient data has required this step. We  
200 averaged substrate class preferences by season (Singh et al. 2010), and we used median values  
201 for species and/or genera for which more than one substrate class or hydraulic preference has  
202 been identified. We excluded taxa for which there were no preferences reported. Because an  
203 exponentially decaying regression model provided the best fit for our raw data, we square-  
204 root transformed our data to make them linear and then performed a linear regression with a  
205 statistical outlier analysis to identify species and genera whose occurrence was most probably  
206 influenced by factors other than substrate type or shear stress. Outlier analysis was performed  
207 using functions in the car statistical package (Fox & Weisberg, 2011) with Cook's distance  
208 and hat-values. Hat-values exceeding threefold the average hat-value, combined with Cook's  
209 distances values  $> 4/n$ , were considered outliers and removed from later analyses. We used  
210 hat-values exceeding half of the abovementioned common thresholds (1.5 times the average  
211 hat-value and Cook's distance values  $> 2/n$ ) to identify species and genera whose occurrences  
212 are most probably determined by environmental factors other than substrate or shear stress. In  
213 so doing, we identified 13 outliers with our stepwise outlier approach (Online Resource 2), all  
214 of which we discuss in our chapter on environmental co-variables determining species  
215 presence.

216 Using piecewise regressions (*piecewise.linear* function, SiZer package) (Sonderegger,  
217 2012), we analyzed the trimmed data set to determine the average shear stress threshold,  
218 which we assumed could be used to define the habitat suitability for the species. Piecewise  
219 regression analysis (also known as segmented regression or broken-stick regression) is a type  
220 of regression analysis where the response variable is partitioned into intervals to which  
221 subsequently separate regressions are fit. This allows one to calculate critical thresholds along  
222 a response curve, which can identify specific thresholds of preferred shear tolerance among  
223 taxa. In addition, we performed a lognormal species sensitivity analysis to determine the  
224 average shear stress preferences of the entire data set using the *fitdistr* function from the  
225 MASS package (Venables & Ripley, 2002). This sensitivity analysis was done to identify the  
226 percentage of species limited by habitat suitability constraints, which we assumed would  
227 occur at the average shear stress preference larger than the critical threshold identified by the  
228 piecewise regression.

229

## 230 **Results**

231

### 232 Variability in hydraulic and substrate preferences

233 Two data sets were compiled for invertebrate species, one comprised of 215 hydraulic shear  
234 stress preferences for 188 invertebrate taxa in 63 families (Online Resource 1); the other data  
235 set was comprised of 209 substrate preferences and 189 sediment grain size preferences  
236 (Online Resource 3). Preferred sediment grain sizes ranged from 0.05 to 400 mm and  
237 preferred shear stress values ranged from 0.13 to 3.67 N m<sup>-2</sup>.

238 Variations in hydraulic preferences did not differ significantly ( $p = 0.09$ ) among  
239 species within the same genus or family (median IQR at genus level = 1.9, median IQR at  
240 family level = 2.7). However, hydraulic preferences may vary among families within the same

241 order, which is illustrated by the example diagrams (whisker plots) of hydraulic preferences  
242 for five major invertebrate groups, pooled by genus or family (Fig. 1). The Wilcoxon rank  
243 sum test revealed a slight trend toward higher median IQR scores at the genus level than at  
244 higher taxonomic levels. As a result, our assessment of preferences at taxonomic levels higher  
245 than families may have grouped together species with significantly different flow  
246 requirements (e.g., Elmidae and Dytiscidae (Fig. 1)). Substrate preferences did not vary  
247 significantly ( $p = 0.51$ ) between species within the same genus or family (median IQR at  
248 genus level = 5, median IQR at family level = 5), due to the large range in IQRs in both  
249 groups.

250

#### 251 Flow resistance thresholds

252 We identified six studies that provided data on tolerated shear stresses for 27 taxa (Online  
253 Resource 4), all of which considered variations in body size (Schnauder et al., 2010; Hauer et  
254 al., 2012) or the habitat wherein the individuals were located at the moment the drift causing  
255 event occurred (Borchardt, 1993; Gabel et al., 2012). For 14 of these 27 taxa, data were  
256 available for both hydraulic preference and critical shear stress (Online Resources 1 and 4).  
257 Although our comparisons show that the dislodging shear stress exceeds the preferred  
258 conditions for most species (as expected), the opposite relationship is seen for the caddisfly  
259 *Rhyacophila* spp. and the mayflies *Epeorus sylvicola* and *Rhithrogena semicoloranta* (Online  
260 Resource 4). This seemingly difference in tolerances might have been due to the responses of  
261 species to other environmental factors that influence their distributions (as illustrated by  
262 *Rhithrogena* sp. in Fig. 2), such as oxygen requirements.

263

#### 264 Relationship between preferred substrate size and shear stress

265 The regression model of substrate preferences plotted against shear stress preferences yielded  
266 preference relationships for 41 taxa and two groups of outliers (Fig. 2). Taxa in the upper part  
267 of the graph are outliers (*Pisidium* sp., *Elmis* sp., the caddisfly *Philopotomus* sp., the mayfly  
268 *Rhithrogena* sp., and the chironomids Tanytarsini gen. sp., Tanypodinae gen. sp. and  
269 *Polypedilum breviantennatum*) that prefer finer substrates than indicated by their shear stress  
270 preferences. In contrast, in the lower part of the graph, taxa such as *Amphinemura* sp., *Perla*  
271 sp. and *Caenis* sp. prefer coarser substrates than their shear stress preferences suggest. The  
272 presence of these outlier taxa may result from environmental factors that can only be provided  
273 under conditions where ecological processes are intact, thus suggesting that the taxa could  
274 serve as sensitive indicator species when evaluating alteration to hydromorphological  
275 integrity (see Discussion).

276

#### 277 Thresholds of physical forces limiting habitat suitability

278 Our piecewise regression model identified a shear stress of  $0.64 \text{ N m}^{-2}$ , which represents the  
279 physical threshold limiting the habitat suitability of species (Fig. 2). This suggests that of the  
280 188 taxa we analyzed, 39% are potentially affected by limited habitat suitability where shear  
281 stresses exceed  $0.64 \text{ N m}^{-2}$  (Fig. 3).

282

#### 283 **Discussion**

284 In general, there is very limited quantitative data available on species' responses to  
285 hydromorphological conditions in streams (Wolter et al., 2015). Of the approximately 20,000  
286 European freshwater invertebrate species described, we found quantitative data on substrate  
287 size preferences for only 189 taxa and hydraulic preferences for 188 taxa. Although this  
288 sample comprises a rather small proportion of all known freshwater invertebrate taxa in  
289 Europe, our data set is still the largest compilation available on the relationship between

290 sediment size and shear stress tolerances for stream invertebrate taxa. This small sample size  
291 is problematic because environmental assessment protocols, such as the PERLODES  
292 assessment system, rely on life history data of approximately 1100 invertebrate taxa (Meier et  
293 al., 2006).

294

### 295 Hydraulic and substrate preferences

296 Our results show that hydraulic and substrate preferences of invertebrate species vary  
297 widely within taxonomic orders, among families within the same order, but not at the family  
298 and genus levels. We found that trichopterans and ephemeropterans, groups that are  
299 frequently targeted in river restoration (Lorenz et al., 2016b), show large interspecific  
300 variations in their hydraulic preferences. However, for for planning restorations, it would be  
301 useful if there were indicator species available that could better diagnose ecological  
302 conditions (Monk et al., 2012).

303 Variations in substrate size have been shown to be a significant predictor of  
304 invertebrate diversity (Waters, 1995; Angradi, 1999; Buss et al., 2004; Jowett, 2003). Our  
305 results show a slight trend of taxa showing higher variation in substrate preferences at the  
306 family level than at the genus level, a difference that might become more pronounced as more  
307 data become available. Erosion and deposition of fine sediments and organic matter influence  
308 the abundance of invertebrates (e.g., the abundance of Ephemeroptera, Plecoptera and  
309 Trichoptera species decline in finer mineral substrates) (Reice, 1980; Maxted et al., 2003;  
310 Beauger et al., 2006; Allan & Castillo, 2007; Timm et al., 2008; Duan et al., 2009; Pan et al.,  
311 2012). These reported finding suggests that higher habitat heterogeneity and complexity are  
312 associated with higher invertebrate diversity (Buss et al., 2004).

313 Although other studies have found only minor effects of habitat improvement on  
314 benthic invertebrate diversity, habitat improvement has been shown to exert large effects on

315 dispersal ability of local species pools (Sundermann et al., 2011b; Tonkin et al., 2014).  
316 However, none of those habitat-improvement studies included detailed information on the  
317 effect of preferred shear stress or substrate availability on diversity, even though substrate  
318 condition is known to affect benthic invertebrate assemblages, especially following stream  
319 restoration (Jähnig & Lorenz, 2008; Lorenz et al., 2009). Ideally, a river restoration should  
320 provide a variety of habitats that provide the suite of desirable grain sizes preferred by  
321 epifaunal species, even though environmental stresses may still exist or even prevail at the  
322 catchment scale (e.g. Sundermann et al., 2013). Based on our data, river restoration designs  
323 could explicitly provide a variety of hydraulic conditions that would generate a variety of  
324 substrate sizes. These substrates could be generated by creating structures in channels, such as  
325 riffles, pools, coarse wood, and river banks (Lange et al., 2015), which could be designed to  
326 meet habitat requirements of specific target species. Our results indicate that if detailed  
327 information on target species is unavailable for restoration planning, then data for families  
328 could be used to select hydraulic and substrate targets.

329

### 330 Flow resistance thresholds

331 Critical flow velocities, which define the thresholds at which invertebrates become detached  
332 and dislodged, are typically much higher than their preferred flow velocities (Söhngen et al.,  
333 2008). However, data shown by Söhngen et al. (2008) indicated that a surprisingly high  
334 number of invertebrate species are able to resist flow velocities close to their detachment  
335 thresholds. The shear stress data we collected had been measured in rivers of various sizes  
336 and using various methods, and this may account for some of the variation we encountered.  
337 However, particle size variations in substrates (at small scales) may also explain why different  
338 species exist in different hydraulic environments. Hydraulic preferences for species adapted to  
339 higher shear stress than their critical shear stress (which induces dislodgement) (e.g., for

340 *Rhyacophila* spp., *E. sylvicola*, and *R. semicoloranta*) were derived from surveys in the large  
341 Upper Rhône River (France) (Mérigoux et al., 2009). However, critical shear stress thresholds  
342 for these species were determined from individuals sampled in the much smaller Ybbs River  
343 (Austria) (Hauer et al., 2012). Apparently, hydraulic preferences of epifaunal species (shear  
344 stress, but also water depth and velocity) vary with river size (Jowett, 2000; Mérigoux et al.,  
345 2009). However, higher critical shear stress thresholds might also be related to the interaction  
346 of increasing flow velocity with higher bed instability [e.g., *Baetis* sp. inhabited conditions  
347 ranging from 0.26 N m<sup>-2</sup> to 9 N m<sup>-2</sup> (Gibbins et al., 2010; Hauer et al., 2012)]. Species  
348 probably benefit when they can access low flow areas (refuges) in interstitial spaces of coarse  
349 substrates. When coarse sediments are washed away, organisms are more likely to detach and  
350 drift downstream.

351         Critical flow velocities and shear stresses needed to dislodge invertebrates also vary  
352 considerably, depending on the body shape and size of the organism. Larger specimens are  
353 typically more likely than smaller individuals and juveniles to actively resisting high  
354 hydraulic forces (Schnauder et al., 2010; Hauer et al., 2012). Additionally, species might  
355 respond differently to a slow and continuous increasing of hydraulic stress by actively  
356 avoiding areas of high flow. For example, some species respond by burrowing deeper into the  
357 sediment or by adjusting body postures to minimize drag when flow conditions gradually  
358 exceed some threshold rate (Poole & Stewart, 1976; Statzner, 1981; Holomuzki & Biggs,  
359 2000; Schnauder et al., 2010). In contrast, abrupt changes in hydraulic conditions are more  
360 likely to dislodge invertebrates than gradual increases of flow velocity, discharge, or shear  
361 stress (Imbert & Perry, 2000).

362         Our results indicate that behavioral responses to high flow conditions are species-  
363 specific and do not occur in all invertebrate species. For example, the damselfly  
364 *Calopteryx splendens* can resist a gradual increase in shear stress of 2.4 N m<sup>-2</sup> in sand by



365 adopting a drag-minimizing body posture (Schnauder et al., 2010), but can be dislodged by an  
366 abrupt change in shear stress of only  $0.13 \text{ N m}^{-2}$  (Gabel et al., 2012) (Online Resource 3).  
367 However, peak shear stress conditions seem to be the primary cause for dislodgement of taxa  
368 that are unable to attach to a sediment surface (Online Resource 4). This interpretation is  
369 supported by the behavior of the gastropod *Bithynia tentaculata*, which is dislodged at shear  
370 stresses between  $0.44$  and  $0.57 \text{ N m}^{-2}$ , whether stress increase is gradual or abrupt (Online  
371 Resource 4). This example suggests that mobile species are able to adapt behaviorally to  
372 gradual increases in flow more so than sessile species. However, other gastropods can quickly  
373 produce mucus to anchor themselves (a behavior not observed for *B. tentaculata*) when shear  
374 stresses increase above a particular threshold (Schnauder et al., 2010). Even after  
375 dislodgement, remaining mucus threads provide an adhesive anchorage that prevents the  
376 snails from drifting very far downstream, suggesting that some gastropod species might also  
377 possess species-specific resistance behaviors (Schnauder et al., 2010) that may affect  
378 hydraulic and substrate preferences.

379

#### 380 Relationship between preferred substrate size and shear stress

381 Our results identified nine taxa not following the estimated shear stress/substrate relationship  
382 in the upper part of Fig. 2. Most of those nine taxa have been well documented as having  
383 extremely high dissolved oxygen requirements. For example, riffle beetles (Elmidae: *Elmis*  
384 sp.) prefer cool, fast-flowing streams and rivers, conditions in which water is almost  
385 completely saturated with oxygen (Elliott, 2008). Likewise, the mayfly *Rhithrogena* sp.  
386 typically occurs in rapidly-flowing, oxygen-rich waters and is sensitive to low dissolved  
387 oxygen concentrations that occur in slowly flowing water (Ambühl, 1960), whereas  
388 chironomid species of tanytarsini (*Micropsectra* sp.) and tanypodinae (*Procladius* sp.) are  
389 sensitive to intermediate dissolved oxygen conditions (Bérg et al., 1962; Johnson, 1995;

390 Quinlan & Smol, 2001; Brodersen et al., 2008). In contrast, sphaeriids of the genus *Pisidium*  
391 sp. are resistant to hypoxia, but they depend on locally available oxygenated microhabitats  
392 within sediments where they can persist during hypoxic events (Mackie, 2007).

393 For stream epifauna species that require highly oxygenated conditions, oxygen concentration  
394 may be more important to them than size of substrate material (Fig. 2). However, our results  
395 show that such species seem to prefer finer substrates than would be expected to occur in high  
396 shear stress environments (Fig. 2). This holds true despite we even had to pool species that  
397 show a large range of preferred grain sizes, e.g. the taxon *Rhithrogena* sp. (Online Resource  
398 3). The available preferred substrate data of this species vary between investigated seasons,  
399 with larger substrate preferences in spring (cobbles) compared to other seasons. Spring is this  
400 season where large amounts of fine sediment are flushed from the stream channel due to rain  
401 events and snow melting. Similarly, this species prefers larger sediments in mountain areas  
402 (where large sediments such as boulders or cobbles are dominating), while it prefers finer  
403 sediment in other areas. This strongly points on the interpretation that the presence of this  
404 species is not primarily driven by hydromorphological constraints, even though substrate  
405 preferences may match hydraulic preferences in certain areas (mountain) or seasons (spring)  
406 of the year. To cope with this trade-off between preference for fine substrates and high  
407 oxygen availability, these species probably seek moderately flushed, but well oxygenated  
408 interstitial spaces that provide them both shelter from high flow conditions and access to fine  
409 substrate.

410 We found that some taxa had been characterized by significantly lower shear stress tolerance  
411 than would have been expected by their preferred substrate type (Fig. 2). For example, the  
412 taxa *Amphinemura* sp. and *Perla* sp. are strong biofilm grazers (Graf et al., 2007; Graf et al.,  
413 2009), which might explain their preference for large substrate size where biofilm tend to  
414 grow. Likewise, although Singh et al. (2010) reported that *Caenis* sp. is associated with large

415 stones and boulders, the species is a detritivore, which means it typically seeks patches of  
416 organic detritus and decaying leaves (Int Panis et al., 1994; Pabst et al., 2008) that typically  
417 accumulate between boulders in slowly to moderately flowing rivers. Therefore, the  
418 preference of *Caenis* for boulders identified by Singh may primarily reflect the species  
419 preference for organic food resources rather than a preference for inhabiting large substrate  
420 (Bradbeer & Savage, 1980). Other taxa showing preferences for unusual or unexpected  
421 hydraulic/substrate combinations might actually be responding more strongly to a particular  
422 resource or oxygen availability than to substrate type or flow condition in a stream (Kakouei  
423 et al., 2017). Such taxa might be useful as sensitive indicator species for evaluating ecological  
424 conditions (and stream integrity) because their presence might indicate sufficient habitat  
425 heterogeneity.

426

#### 427 Thresholds of physical forces limiting habitat suitability

428 It is well known that changes in hydrological conditions in streams can directly affect habitat  
429 quality for epibenthic invertebrates (Sousa, 1984; Suren & Jowett, 2006; Dewson et al.,  
430 2007). When stream waters exceed a certain flow rate, the higher power results in a more  
431 homogenized sediment condition and low accumulations of fine organic materials (leaves and  
432 detritus) in sheltered areas (Shvidchenko et al., 2001). Similarly, sedimentation increases with  
433 slow flow and the accumulation of fine sediments leads to homogenized sediment conditions  
434 as well. Such lowering of habitat complexity may consequently exclude taxa that depend on  
435 specific environmental conditions that are no longer available (e.g., those with fine substrate  
436 requirements or specific feeding behaviors (Graeber et al., 2013; Lorenz et al., 2016b; Stoll et  
437 al. 2016)). Therefore, river rehabilitation planning should seek to create a diverse and  
438 heterogeneous flow environment (Kakouei et al., 2017), which is expected to generate more  
439 complex habitat conditions (Lorenz et al., 2016a).

440 River rehabilitation projects typically enhance physical and hydraulic habitat  
441 complexity to increase in-stream biodiversity (Milner & Gilvear, 2012). However, for 78  
442 restoration projects reviewed by Palmer et al. (2010), only two resulted in an increase in  
443 invertebrate diversity. Palmer et al. (2010) attributed the failure to meet species diversity  
444 targets to the focusing of the project almost exclusively on improving habitat structure to  
445 increase species diversity at the expense of improving other, potentially more important  
446 environmental conditions (Lake, 2000; McCabe & Gotelli, 2000; Ward & Tockner, 2001;  
447 Menninger & Palmer, 2006; Muotka & Syrjänen, 2007; Warfe et al., 2008). The restorations  
448 that focused exclusively on habitat might have failed because suitable sediment sizes (a  
449 reflection of promoting an appropriate shear stress) may not have been created, especially if  
450 the underlying hydromorphological processes were not addressed. The hydraulic threshold for  
451 biota we estimated in this study might be helpful in guiding future rehabilitation planning  
452 focused on improving in-stream biodiversity, even though several species we examined may  
453 have hydraulic preferences exceeding  $0.64 \text{ N m}^{-2}$ .

454

#### 455 Rehabilitation implications

456 Our synthesis was focused on identifying the response of stream epibenthic biota  
457 (invertebrates) to hydromorphological and physical forces that might limit habitat suitability.  
458 Therefore, our main objective was to compile all available data on invertebrate habitat  
459 tolerances relative to shear stress and substrate size to test the relationship between preferred  
460 shear stresses and preferred sediment sizes.

461 River rehabilitation must take into account a myriad of potential competing factors,  
462 such as dredging and channelization for navigation, erosion control, and water extraction, all  
463 of which typically hamper hydromorphological processes in streams (sediment erosion,  
464 sediment transport downstream, and sediment sorting, large wood accumulation). In addition,

465 water managers and rehabilitation planners should consider flood protection and bank stability  
466 when they identify rehabilitation targets, both of which can constrain possible restoration  
467 targets. Under such constraints, successful river rehabilitation will benefit from precise  
468 information on preferred and tolerated sediment or substrate size distributions and shear stress  
469 thresholds so that restorations can both incorporate safety requirements into the plan and  
470 optimize ecological improvements.

471         We found that riverine benthic invertebrates inhabit a wide range of substrate sizes  
472 (0.05–400 mm) and hydraulic conditions (shear stresses of 0.13–3.67 N m<sup>-2</sup>). This suggests  
473 that to maximize biodiversity, a wide range of physical and hydraulic habitat heterogeneity  
474 can be considered when planning and designing instream rehabilitation projects. Although  
475 hydraulic and substrate preferences did not significantly vary among species within the same  
476 genus or family, impact assessment and restoration planning should seek to define appropriate  
477 indicator taxa at the species level whenever possible. This is because using higher taxonomic  
478 groups may incorporate too much variability for defining rehabilitation targets. That is, the  
479 member species may have opposing requirements and/or flow thresholds and we do not yet  
480 know the habitat requirements of all those species. However, if detailed information on  
481 habitat preferences of some invertebrate species is unavailable, using currently known  
482 preferences at the family scale can be used to define restoration target conditions or evaluate  
483 restoration success.

484         Physical and hydraulic habitat heterogeneity and complexity are commonly viewed as  
485 the main factors structuring invertebrate communities. Our synthesis of the literature  
486 demonstrated that a part of the stream invertebrate fauna may be affected to a higher extent by  
487 additional environmental factors, such as dissolved oxygen availability and the  
488 presence/absence of food resources, than by physical and hydraulic constraints. Therefore,  
489 many of the weak biotic responses reported in the restoration literature may have been

490 resulted from ecological stresses (e.g., oxygen concentrations or food resources) not being  
491 adequately addressed, stresses that are independent of creating adequate physical  
492 hydromorphological conditions (e.g., habitat structure) (Palmer et al., 2010; Bernhardt et al.,  
493 2005; Barnes et al., 2013; Haase et al., 2013).

494

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847 ***Electronic Supporting Information***

848 Additional supporting information may be found in the online version of this article:

849

850 Online Resource 1. Hydraulic preferences of benthic invertebrates.

851 Online Resource 2. Hat values and Cook's distance calculated from the linear regression

852 models on squareroot-transformed data on substrate vs. hydraulic preferences of benthic

853 invertebrates.

854 Online Resource 3. Substrate and grain size preferences of benthic invertebrates.

855 Online Resource 4. Critical shear stress thresholds detaching and dislodging invertebrates.

856

857

858 *Tables*

859 Table 1: Sediment grain sizes of substrate classes examined in this study.

860

Substrate class	Sediment grain size [mm]
0	$256 \leq 128$
1	$128 \leq 64$
2	$64 \leq 32$
3	$32 \leq 16$
4	$16 \leq 8$
5	$8 \leq 4$
6	$4 \leq 2$
7	$2 \leq 1$
8	$1 \leq 0.5$
9	$0.5 \leq 0.25$
10	$0.25 \leq 0.125$
11	$0.125 \leq 0.05$

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862

863 ***Figure captions***

864

865 Fig. 1: Hydraulic preferences of epibenthic invertebrates (Coleoptera, Diptera, Plecoptera,  
866 Trichoptera, and Ephemeroptera), pooled by genus or family. The taxonomic identity for each  
867 group are provided as examples and do not represent any qualitative or quantitative ranking  
868 within the group. Whisker plots represent mean  $\pm$  1 standard deviation.

869

870 Fig. 2: Regression model of substrate preferences of benthic invertebrates from class 0 to 11  
871 plotted against their hydraulic preferences (solid circles). Outliers (open circles) indicate taxa  
872 not primarily determined by hydromorphologic conditions. Solid red line shows results of the  
873 piecewise regression model indicating the significant inflection point (change point) where  
874 shear stress limits habitat suitability. The dashed line shows the linear regression model  
875 following the substrate size/shear stress preference relationship, excluding outliers (adj.  $r^2 =$   
876 0.36,  $p = 0.002$ ). The dashed line at the top of graph marks the threshold between substrate  
877 class 11 and detritus (Det.) as preferred substrate.

878

879 Fig. 3: Species sensitivity distribution of benthic invertebrate taxa against hydraulic  
880 preference ( $N\ m^{-2}$ ). Different colors indicate different taxonomic groups. Solid red line  
881 indicates fitted lognormal distribution and dashed lines indicate 95% confidence intervals.  
882 Solid grey lines indicate the shear stress threshold where 39% of the 188 taxa analyzed  
883 become potentially affected by limited habitat suitability (i.e., when the threshold of  $0.64\ N$   
884  $m^{-2}$  is exceeded).

885

886